

Removal of Hg(II) and acid yellow dye using chitosanmontmorillonite nanocomposite

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Abstract: Chitosan-glutaraldehyde / nano-montmorillonite composite has been successfully prepared. The structure of the prepared bio sorbent has been confirmed by Fourier transforms infrared spectroscopy (FTIR), transmission electron microscopy (TEM) and scanning electron microscopy (SEM). The adsorption kinetics and isotherm behaviors of Hg(II) and acid yellow dye (A17) molecules onto the prepared chitosan nanocomposite had been studied and described by pseudo-first, pseudo-second order and Langmuir, Freundlich isotherm models, respectively. Experimental data exhibit high adsorption capacities equal 125 and 302 mg /g for Hg(II) and A17, respectively.

keywords: Chitosan, Nano-montmorillonite, Adsorption, Hg(II), Textile acid yellow dye.

1. Introduction

There is no doubt that water crisis is the main world issue which attracts public attention due to its harmful effect on respiratory system (i.e., dyspnea), cardiovascular, nervous system of human beings' skin (dermatitis), and eyes (irritation to eyes). Unfortunately, this issue continues to get worse with the use of different synthetic chemicals for various applications including heavy metals and dyes [1-3].

Among heavy metals, Hg(II) has been found to be one of the most toxic metals as it threatens not only human but also aquatic system even though direct exposure [4,5]. Textile dyes such as acid yellow dye 17 (A17) is another severe threat because discharging wastewater down right soil, leaves water unsuitable due to water colorization [6-8].

Several methods such as flocculation, adsorption, biodegradation, chemical oxidation, in addition to photocatalysis have been used for treatment of polluted water. Amongst them, adsorption is an interesting technique owing to its simplicity, effectiveness, economical and greatly efficient [9-11].

Chitosan (CS), as an adsorbent, has been used on a large scale due to possessing a large number of functional groups, abundance, hydrophilicity, biodegradability, and non-toxicity [12,13]. In order to improve its capacity and selectivity, several materials can

incorporate or be grafted with it such as montmorillonite (MMT) clay. It is the mainly utilized adsorbent due to its high active surface area and great cation exchange capacity [14,15]. Therefore, the modification of CS with MMT can enhance the adsorption process towards polluted water [16,17].

Hence, in this study, the potential of chitosan polymer functionalized with nano-montmorillonite as a strong sorbent towards straightforward and rapid removal of Hg(II) ions and A17 was examined.

2. Experimental

2.1. Materials

All chemical substances including chitosan (CS) of 92% deacetylation percentage, montmorillonite (MMT), glutaraldehyde 50%, mercuric chloride (HgCl₂) were provided by Sigma-Aldrich Trading Co., Ltd. In addition, acid yellow dye (A17) was purchased from the Dye Star Company, Brazil.

2.2. Characterization

The chemical structure of the dye has been shown in Fig. (1). FTIR (Perkin Elmer Spectrum BX) absorption spectra of the prepared composite was calculated over a range that varies between 400–4000 cm⁻¹. A Metrohm 692 pH meter was used for pH measurements. A17 concentration was investigated utilizing a Jasco UV - Visible spectrophotometer (model

V-530, Japan) at 425 nm wavelength. The uptake quantity of Hg (II) was studied with the aid of Agilent's 5100 ICP-OES (Agilent technologies, Australia).

2.3. Preparation of chitosan composite

1g of CS was dispersed within 1% aqueous acetic solution and stirring. Then was further displaced into flask for re-precipitation using 0.1M sodium hydroxide solution. Clay (1% w/w of CS) diffused for 24h in water was mixed with the CS solution, agitated with stirrer. Then, NaOH solution was added dropwise under high stirring condition for precipitation. Later, (5% w/w) of glutaraldehyde crosslinking were added dropwise and allowed for cross linking at 45°C. The obtained precipitate was allowed for washing with deionized water then for drying [18]. The expected chemical structure of chitosan nano-montmorillonite composite is shown in Fig.(2).

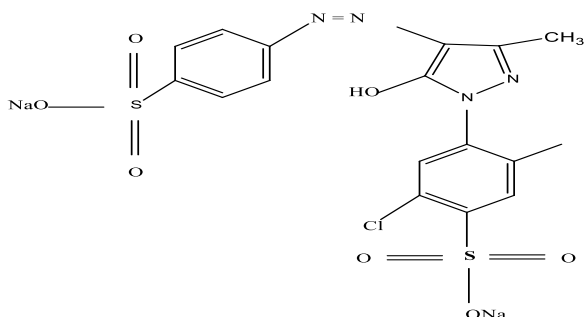


Fig.(1): Chemical structure of acid yellow dye 17 (A17)

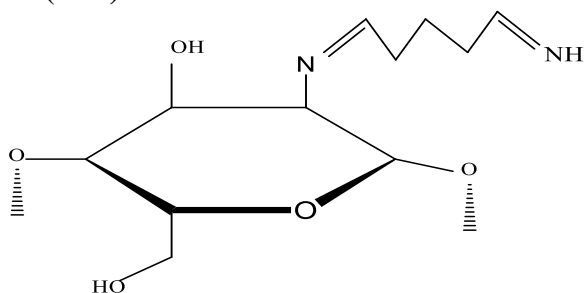


Fig. (2): Proposed chemical structure of CS-nano MMT sorbent

Adsorption studies using batch method.

Adsorption experiments were conducted at 25°C by vibrating 0.02 g of prepared composite alongside 25 mL of initial conc (100 and 20mg/ L) of A17 as well as Hg(II) ranging pH values (2-6) with different time intervals (5-240) min. Uptake value (q_e) was performed using (Eq. 1)

$$q_e = (C_i - C_f) V/W. (1)$$

where C_i and C_f are the initial and equilibrium concentration, respectively. Where V is the solution volume; W is the sorbent mass.

3. Results and Discussion

3.1. Assignment of polymeric sample

3.1.1. FTIR

FTIR investigation of prepared composites, as seen in Fig. (3), the basic absorption peaks of CS (NH stretching, OH stretching, and the intermolecular hydrogen bonds) appeared at 3455 cm^{-1} , bands at 2923 and 2860 cm^{-1} are corresponding to (CH stretching). Peaks at 1652 as well as 1558 cm^{-1} are related to amide (I and II), consequently [18]. Furthermore, MMT shows characteristic peaks at 640 and 464 cm^{-1} suggesting the existence of (Al, Mg)-OH and Si-Al-O bonds, while band near 1048 cm^{-1} is corresponded to Si-O vibrations [19-21]. Afterwards adsorption, the strong band intensity assigned to -OH groups decreased and shifted from 3455 to 3492 cm^{-1} and 3469 cm^{-1} for Hg(II) besides A17, respectively, which obviously confirm that hydroxyl groups have great impact on removing efficiency. Furthermore, the intensity of (-NH) group peak is decreased. As for the prepared composite after adsorption of Hg(II), spikes at 1652 and 1558 are moved to 1641 in addition to 1569 cm^{-1} after adsorption of A17 [22]

cm^{-1} are changed near 1645 and 1554 cm^{-1} respectively, while those peaks. These results indicate the chelation of Hg(II) with CS nitrogen atoms as well as electrostatic attraction between anionic dye and NH_3^+ of CS.

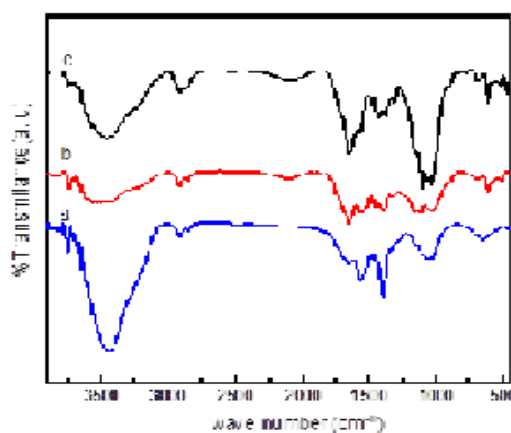


Fig. (3): FTIR spectra for (a) chitosan nano-montmorillonite composite, (b) composite after adsorption of Hg(II), and (c) after A17 adsorption .

3.1.2 TEM

Fig. 4 implies TEM image of CS-MMT nanocomposite in which dark lines refer to nano-MMT while bright ones denote intercalated CS matrix. It is observed that size of particles is lower than 100nm so it suggests well preparation of nanocomposite.

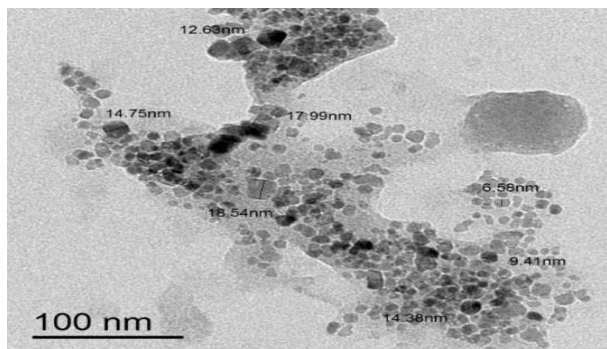


Fig. (4) TEM image of CS-MMT nanocomposite

The surface morphology of prepared nanocomposite after adsorption of A17 and Hg(II), respectively presented in (Fig. 5a and b) display smoother and river-like morphologies.

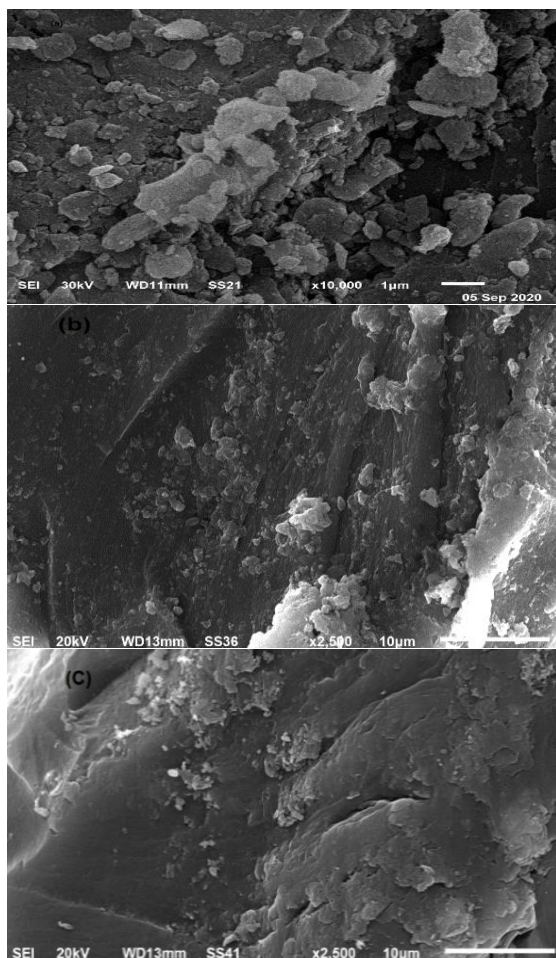


Fig.(5) (a) SEM of CS-MMT, (b) SEM of CS-MMT-A17 and (c)SEM of CS-MMT-Hg(II)

2. Optimization of adsorption of Hg(II) and A17 by CS-MMT nanocomposite

3.2.1. Effect of initial pH

The impact of pH on Hg(II) and A17 adsorption has been studied by stirring of 25 mL mixture of Hg(20 mg/L), (100 mg/ L) of A17 and 20 mg of adsorbent at pH values (2-6). The equilibrium adsorption (q_e) was plotted against the pH. From Fig. (6), it can be concluded that q_e reaches its maximum near pH value equal 4 for both A17 and Hg(II) [23].

3.2.2. Effect of contact time

In order to study the influence of the optimal time on the removal capacity of Hg(II) and A17, experiments were carried on 20 mg of adsorbent, 25 °C, initial concentrations of 20 and 100 mg/ L of Hg(II) and A17, respectively, with each optimal pH value in 25 ml solution at time intervals varying between (5-240) min. Fig.(7) shows that, at the beginning, the adsorption rate increases then remains constant at 60 and 240 min for Hg(II) and A17, respectively. Therefore, 60 and 240 min were chosen as the equilibrium time for the saturation of both metal ion and dye on the prepared adsorbent.

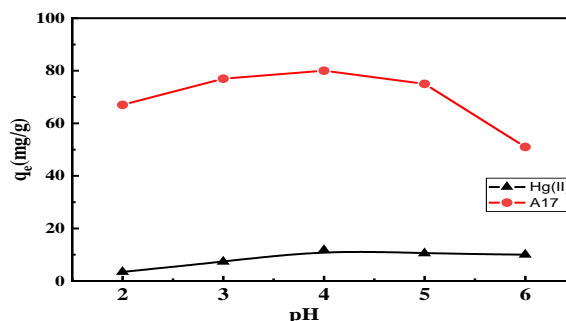


Fig. (6): Influence of pH on of Hg(II) and A17 adsorption onto the prepared composite. 20mg of adsorbent; time= 240 min; temperature=25 °C; C_i (20 and 100 mg/L) for Hg(II) and A17, consequently.

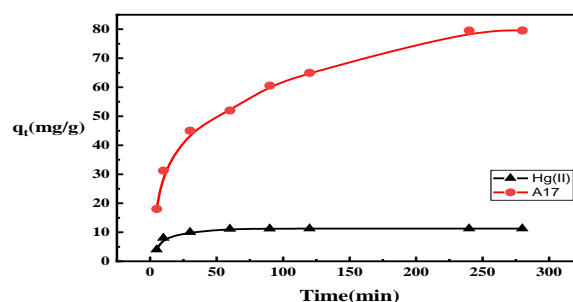


Fig. (7): Influence of equilibrium time on adsorption of Hg(II) and A17 on the prepared composite. 20mg of adsorbent; pH=4;

temperature = 25 °C; C_i (20 and 100 mg/L) for Hg(II) and A17, consequently

3.2.3. Kinetics Analysis

In the kinetics studies for the removal process, the initial pH was adjusted to be 4 for the studied metal and dye carrying out concentrations of 20 as well as 100 mg/L for Hg(II) and A17 near to 25°C. at various time intervals. The fitting of pseudo-first order expressed by (Eq. 2) as well as pseudo-second order (Eq. 3) models were applied to explain the adsorption kinetics [24,25].

$$\ln(q_e - q_t) = \ln q_e - K_1 t \quad (2)$$

$$t/q_t = 1/(k_2 q_e^2) + (1/q_e) t \quad (3)$$

Here, q_t (mg g⁻¹) represents quantity of A17 and Hg(II) adsorbed at (t) time, k_1 (min⁻¹) and k_2 (g mg⁻¹ min⁻¹) are pseudo-first and second order kinetics rate constants, consequently, q_e is the Hg(II) and A17 amount adsorbed at equilibrium.

Fig. (8 a and b) show the experimental curves of pseudo-first and pseudo-second order adsorption kinetic models, respectively, k_1 and k_2 constant values are given in Table (1). Based on data, it has been clarified that model of

Table (1): Constant parameters of pseudo-first-order and pseudo-second-order for adsorption of Hg(II) and A17.

	Pseudo-first-order				Pseudo-second-order		
	$q_{e, \text{exp}}$	$q_{e, \text{cal}}$	k_1	R^2	$q_{e, \text{cal}}$	k_2	R^2
Hg(I)	10.312	2.809	0.0365	0.833	11.031	0.0123	0.999
A17	79.54	42.83	0.0072	0.93625	83.23	0.00036	0.994

Table (2): constants of isotherm model of Langmuir and Freundlich.

	Langmuir isotherm			Freundlich isotherm	
	q_m	k_l	R^2	n	k_f
Hg(I)	400	0.002	0.539	0.9404	0.8269
A17	386.1	0.005	0.886	1.806	2.68

Adsorption isotherm models

In this study, there are two adsorption isotherm models, Langmuir and Freundlich which were performed to evaluate mechanism of adsorption for Hg(II) and A17 dye onto the prepared adsorbent. Langmuir isotherm model,

expressed by Fig.(9a) hypothesises formation of monolayer adsorption on the sorbent surface with fixed active sites and given as illustrated in (Eq.4).

$$C_e/q_e = C_e/q_m + 1/q_m \quad (4)$$

pseudo-second order is the most probable model for estimating sorption kinetics of both Hg(II) and A17, thanks to the highest values of R^2 ($R^2 > 0.99$) besides the well matching of q_e calculated value obtained from pseudo-second-order model and $q_{e, \text{exp}}$ than that provided by pseudo-first-order model.

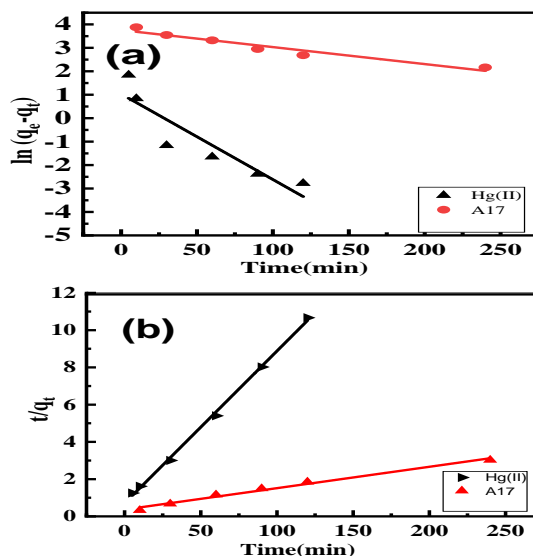


Fig. (8): (a) Pseudo-1st-order, (b) Pseudo-2nd-order kinetics for Hg(II) in addition to A17 adsorption

Hence, q_e is the pollutant amount adsorbed at equilibrium as (mg/ g), q_m (mg /g) is the maximum adsorption capacity of composite, and K_L (L/ mg) is the Langmuir constant. Freundlich isotherm model [26], expressed by Fig.(9b) proposes that adsorption is done on a heterogeneous surface and is explained by the consecutive equation (Eq.5).

$$\text{Log } q_e = \text{log } K_F + 1/n \text{ Log } C_e \quad (5)$$

K_F and n represent Freundlich model coefficients (mg/g). The calculated constant values of the two measured models (Table 2) confirm that Freundlich model predicts well

adsorption mechanism for A17 and Hg(II) on the obtained nanocomposite

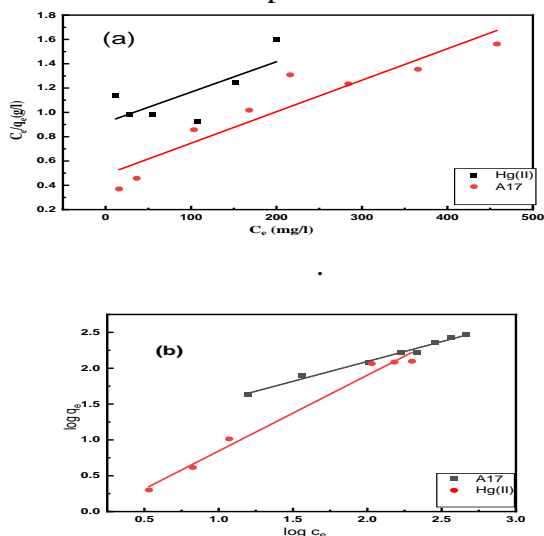


Fig. (9): (a) Langmuir isotherm model, (b) Freundlich isotherm model for adsorption of Hg(II) and A17.

4. Conclusion

The modification of chitosan by montmorillonite as well as crosslinking with glutaraldehyde/ NaOH for using as adsorbent of Hg(II) and A17 was investigated. The prepared composite has been characterized by FT-IR, SEM and TEM, which confirmed a successful grafting reaction. Kinetic studies were depicted better with pseudo-2nd-order kinetic. Freundlich isotherm model described system equilibrium

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