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Physiological Biomarkers of Silver Nanoparticle Toxicity in the Nile Tilapia (Oreochromis niloticus)

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ABSTRACT

Despite their increasing use across various industries, concerns remain about the environmental and health risks associated with nanoparticles (NPs). Among these, silver nanoparticles (AgNPs) are widely applied for their antifungal and antibacterial properties and are incorporated into numerous consumer products. However, limited knowledge exists regarding their toxicity in aquatic environments, prompting the current study. This study investigated the effects of AgNPs on the Nile tilapia (*Oreochromis niloticus*), exposed to concentrations of 0, 1, 3, 6, and 12 μ g/L of AgNPs dissolved in aquarium water for two weeks. Hematological and biochemical parameters, histological examinations, and quantitative real-time PCR (qRT-PCR) analyses were performed. Results indicated dose-dependent effects on immune and blood cell parameters. Immune cell activation—particularly increased levels of neutrophils and monocytes—was observed. At 6 µg/L, red blood cell (RBC) count and hematocrit levels increased, while hemoglobin levels remained relatively stable, with a slight increase at the same concentration. Blood glucose remained relatively stable at lower concentrations (1, 3, and 6 µg/L) but dropped significantly at 12 µg/L. Triglyceride levels increased significantly with AgNP exposure, whereas cholesterol levels decreased significantly at higher concentrations. Among liver enzymes, alanine aminotransferase (ALT) levels showed a significant increase from 3 µg/L onwards, while aspartate aminotransferase (AST) levels increased significantly even at 1 µg/L. Furthermore, a progressive increase in IL-1ß gene expression was observed with increasing AgNP concentrations, peaking at 12 µg/L. In conclusion, AgNPs can be considered safe at lower concentrations, particularly up to 6 µg/L, but caution is advised at higher levels due to their adverse hematological, biochemical, and immunological effects.

INTRODUCTION

Numerous types of nanoparticles (NPs) are now extensively used worldwide (Yan et al., 2012; Yalsuyi et al., 2017). Defined as particles with diameters ranging from 1 to 100 nanometers, NPs possess unique physicochemical properties that distinguish them from their bulk counterparts (Blaise et al., 2008). Despite their expanding applications







across various industries, concerns over the environmental and health-related risks associated with NPs continue to grow (Asharani et al., 2008; Yalsuyi and Vajargah, 2017). As NP production increases, so too does the potential for environmental exposure, particularly in aquatic ecosystems where they may cause harm to aquatic organisms (Klaine et al., 2008; Choi et al., 2010). This has given rise to the emerging field of aquatic nanoecotoxicology, which focuses on studying the ecological impacts of nanoparticles on aquatic environments.

Among the most widely used nanoparticles are silver nanoparticles (AgNPs), which are incorporated into approximately 300 commercial products, including pharmaceuticals, textiles, and inks, due to their well-documented antibacterial and antifungal properties (Mohsenpour et al., 2020). The production and use of AgNPs have grown rapidly, with a projected increase of up to 63% by 2024 (Inshakova, 2017). While their antimicrobial activity has been the subject of significant research (Vali et al., 2020; Mahmoudabadi et al., 2021; Rashidian et al., 2021), their release into water systems—particularly through activities such as washing—raises concerns about environmental contamination.

For over a decade, insufficient data on the toxicity of AgNPs to aquatic organisms has been a major concern (Ale et al., 2018a). Concentrations of AgNPs in aquatic systems have been reported to range from micrograms to milligrams per liter (Gottschalk et al., 2013). Their small size and surface reactivity contribute significantly to their toxicity and persistence in aquatic environments (Ji et al., 2007; Ibrahim, 2020). In water, silver may exist in four oxidation states—Ag⁰, Ag⁺, Ag²⁺, and Ag³⁺—but most studies attribute the toxic effects primarily to the release of Ag⁺ ions, which can be readily absorbed by aquatic organisms (Drake & Hazelwood, 2005; Cáceres-Vélez et al., 2019; Ibrahim, 2020). Even at low concentrations, ionic silver has been shown to harm bacteria and negatively affect freshwater and marine fish (Waalewijn-Kool et al., 2014).

The external fertilization mechanism of fish makes them especially vulnerable to waterborne contaminants such as AgNPs, which may interfere with reproductive success (Johari, 2014). Several studies have demonstrated that metal-based NPs, including AgNPs, can accumulate in fish organs, leading to oxidative stress, impaired physiological functions, and reproductive disruption (Savorelli et al., 2017; Gobi et al., 2018; Ninan et al., 2020). Documented toxic effects of AgNPs include adverse outcomes in species such as the Nile tilapia (*Oreochromis niloticus*), rainbow trout, and common carp (Johari, 2014; Kakavand et al., 2020; Vali et al., 2020).

Tilapia, belonging to the freshwater fish family *Cichlidae*, are native to Africa but have been introduced to many tropical, subtropical, and temperate regions since the mid-20th century (**El-Sayed & Fitzsimmons, 2023**). Today, tilapia are cultivated in over 120

countries and represent the second most important group of farmed bony fish, surpassed only by carp. In 2022, global tilapia production reached 6.55 million metric tons, valued at over 12 billion USD, with the Nile tilapia (*O. niloticus*) accounting for 5.3 million metric tons of that total (**FAO**, **2024**). This species alone constitutes more than 80% of global farmed tilapia production.

The Nile tilapia are known for their resilience to harsh environmental conditions, including polluted waters, making them effective bioindicators for aquatic contaminants (Ma et al., 2018). Given their ecological and economic importance, this study aimed to investigate the chronic toxicity of AgNPs on the hematological and biochemical profiles of O. niloticus during a 14-day exposure period. Plasma biochemical parameters serve as reliable biomarkers for fish health (Ramírez-García et al., 2018; Burgos-Aceves et al., 2019), while hematological analysis offers a non-lethal and efficient method for evaluating general health status in fish (Kumar et al., 2022).

MATERIALS AND METHODS

Characterization of colloidal AgNPs

Plasmonic silver nanoparticles (AgNPs) used in this study were obtained from NanoTech Egypt for Photo-Electronics Communication Center (Brand: NT-SNP).

- Size and Shape: Transmission electron microscopy (TEM) was conducted using a JEOL JEM-2100 high-resolution TEM operating at 200 kV to determine the size and morphology. The AgNPs had an average diameter of 45 ± 5 nm and exhibited a spherical-like shape.
- Optical Properties: UV-Vis absorption spectra were recorded using an Ocean Optics USB2000+VIS-NIR fiber optic spectrophotometer.
- General Characteristics: The AgNPs were yellow in appearance, water-soluble, and provided as a colloidal solution.

Fish and experimental facility

The experiment was conducted at the Fishery Physiology Laboratory, National Institute of Oceanography and Fisheries (NIOF), Alexandria, Egypt. Juvenile Nile tilapia (*Oreochromis niloticus*), with an average initial weight of 15.08 ± 0.88 g, were obtained from the El-Mex Experimental Station of NIOF.

Fish were randomly distributed into triplicate groups, with eight fish per 12-liter glass tank, under a 12 h light:12 h dark photoperiod. All tanks received a continuous

supply of aerated, dechlorinated tap water maintained at 25 ± 1 °C. Fish were fed a commercial diet containing 30% crude protein at 3% of body weight daily, and feeding was suspended 24 hours before and during the experiment.

Experimental treatments included exposure to 0, 1, 3, 6, and 12 μ g/L of AgNPs, directly dissolved in the aquarium water. Each tank served as an independent experimental unit. Fish were fed to apparent satiation at 09:00, 12:00, and 15:00 daily. Fecal matter was removed by daily siphoning. The exposure period lasted 14 days.

All procedures were conducted in compliance with the Institutional Animal Care and Use Committee (IACUC) guidelines and approved by the Alexandria University Ethics Committee (Approval No. AU04 601 25 11 2024).

Blood and tissue sample collection

At the end of the exposure period, three groups of eight fish were sampled from each treatment. Control fish were handled identically to treated groups to ensure uniform stress conditions. Fish were captured quickly using a hand net to minimize stress. Blood samples were collected via cardiac puncture using a heparinized glass pipette.

Collected blood was divided for hemoglobin and plasma biochemical analysis. Plasma was immediately separated by centrifugation and stored at $-20\,^{\circ}$ C to prevent hemolysis. Following blood collection, fish were euthanized, and liver and gill tissues were dissected and preserved for histological analysis.

Hematological and biochemical analysis

- Hemoglobin concentration was measured using a commercial Diamond Diagnostic Kit.
- Plasma glucose was measured using the GOD-PAP enzymatic method.
- Red blood cell (RBC) and white blood cell (WBC) counts were performed according to **Schaperclaus** *et al.* (1991). For RBC, 20 µl of blood was diluted in 3980 µl of RBC diluting solution and counted using a Neubauer Hemocytometer (Rohem, India).
- Hematological indices including mean corpuscular hemoglobin (MCH), mean corpuscular hemoglobin concentration (MCHC), and mean corpuscular volume (MCV) were calculated as per **Haney** *et al.* (1992).
- Plasma cholesterol and triglyceride levels were measured using CHOD/POD and GPO/PAP method kits, respectively.

• Liver enzyme activity of aspartate aminotransferase (AST) and alanine aminotransferase (ALT) was assessed following the manufacturer's protocol (Biolabo, France), using a Mispa CCXL clinical chemistry analyzer (AGAPPE, Switzerland).

Histological investigation

Liver and intestinal tissues were collected from each treatment group and fixed in 10% neutral-buffered formalin. The samples were dehydrated using a graded ethanol series (70% to 100%), cleared with xylene, and embedded in paraffin. Serial sections (5 µm thick) were prepared using a Leica RM 2155 microtome (England) and stained with hematoxylin and eosin (H&E) according to **El-Sayed** *et al.* (2014).

RNA extraction and quantitative real-time PCR (qRT-PCR)

Hepatopancreas samples were excised on ice using a sterile scalpel, snap-frozen in liquid nitrogen, and stored at -80 °C for gene expression analysis.

- RNA extraction was performed using the GenElute Mammalian Total RNA Miniprep Kit (Sigma-Aldrich, USA).
- RNA concentration and purity were assessed using a NanoDrop One spectrophotometer (Thermo Fisher Scientific, USA).
- cDNA synthesis was conducted using PrimeScript RT Master Mix (Thermo Fisher, USA).
- qRT-PCR was performed in triplicate using the QuantStudio 7 Pro Real-Time PCR System (Applied Biosystems, USA) with SYBR Premix Ex Taq (Thermo Fisher, USA).

 β -actin was used as the housekeeping gene and GAPDH as the internal reference. Primer sequences and NCBI GenBank accession numbers for the target gene Interleukin-1 β (IL-1 β) and reference genes are listed in Table (1)

Table 1. Primer sequence and accession number used for q-PCR analysis

GENE NAME	ACCESSION NO	FORWARD SEQUENCE >3')	(5'-	REVERSE SEQUENCE (5'->3')	PRIMER LENGTH
INTERLEUK IN-1 BETA (IL-1B)	XM_019365844.2	GACACTGCTTCTGAAC ACAAGT	CT	TCAGCACTGGCTCTG AAGTG	209

Statistical analysis

Statistical analysis was performed using SPSS software (version 20.0 for Windows; SPSS Inc., USA). Data normality was assessed using the Shapiro–Wilk test. One-way analysis of variance (ANOVA) was employed to evaluate differences among treatment groups, with statistical significance determined at P < 0.05 based on 95% confidence intervals. *Post hoc* comparisons were conducted using Tukey's test. Control fish were exposed to the same level of disturbance as treated fish, including transfer to a different aquarium, while remaining in their original water conditions. Results presented in tables are expressed as mean \pm standard error of the mean (SEM).

RESULTS

Hematological parameters (complete blood count "CBC")

The results showed that AgNPs exposure has dose-dependent effects on immune and blood cell parameters (Table 2). At lower concentrations (1 μ g/l), activation of immune cells (neutrophils and monocytes) was observed, and at higher concentrations (6 μ g/l), RBC and hematocrit levels increased. Conversely, hemoglobin levels remained mostly stable, with an insignificant increase at 6 μ g/l.

Table 2. Complete blood count (CBC) of the Nile tilapia (*Oreochromis niloticus*) exposed to AgNPs for 2 weeks (mean \pm SE; n = 5)

PARAMETER	CONTROL	1 μG/L AGNPS	3 μG/L AGNPS	6 μG/L AGNPS	12 μG/L AGNPS
LYMPHOCYTES (%)	44 ± 2	25 ± 3 a, c, d, e	41 ± 2 a, b, d, e	39 ± 1 a, b, c, e	45 ± 2 b, c, d
NEUTROPHILS (%)	53 ± 3	$65 \pm 4^{a,c,d,e}$	$53 \pm 2^{b,e}$	53 ± 1 ^{b, e}	$50 \pm 2^{a,b,c,d}$
MONOCYTES (%)	2 ± 1.06	$8 \pm 0.90^{a,c,d,e}$	$5 \pm 0.30^{a,b}$	$6 \pm 0.30^{\text{ a, b, e}}$	$4 \pm 0.60^{a, b, d}$
EOSINOPHILS (%)	1 ± 1.17	$2 \pm 0.30^{a,c,e}$	$1 \pm 0.61^{\text{ b, d}}$	$2 \pm 0.69^{\text{ a, c, e}}$	$1\pm0.67^{\text{ b, d}}$
WBCS (10 ³ /CMM)	92.8 ± 1.3	$90.13 \pm 0.5^{\text{ a, c, d,}}$	$115.13 \pm 0.9^{\mathrm{a,b,d,e}}$	$98.1 \pm 2.0^{a,b,c,e}$	$94.2 \pm 0.9^{\text{ a, b, c,}}$
RBCS (10 ⁶ /CMM)	1.15 ± 0.05	$1.00 \pm 0.1^{\text{ a, c, d,}}$	$1.20 \pm 0.18^{b, d}$	$1.41 \pm 0.25^{a, b, c, e}$	$1.16 \pm 0.20^{\text{ b, d}}$

HB (G DL ⁻¹)	5.9 ± 0.82	5.9 ± 0.05 c, d, e	$6.1 \pm 0.80^{\text{ a, b, d}}$	$7.2 \pm 0.33^{\text{ a, b, c, e}}$	6.0 ± 1.11 b, d
PLATELETS (10 ⁹ /CMM)	110 ± 0.9	$110\pm0.5~^{c,d,e}$	$118\pm0.3^{a,b,d,e}$	115± 0.5 ^{a, b, c, e}	112± 0.4 a, b, c, d
MCV(%)	159± 0.6	$156 \pm 0.8^{c,d,e}$	150± 0.9 a, b, d, e	158.2± 0.5 a, b, c	158.2± 0.4 a, b, c
MCH (%)	51.3± 0.6	$49 \pm 0.8^{a,c,d,e}$	$48.6\pm0.6^{a,b,d,e}$	51.1± 0.5 b, c	50.9± 0.4 ^{a, b, c}
HAEMATOCRIT (%)	18.3± 0.42	$18.6 \pm 0.81^{c, d, e}$	15.8± 0.92 a, b, d	22.3± 0.62 ^{a, b, c,}	17.4± 0.95 b, d

Data are expressed as mean \pm SEM. n = 8 for each experimental group

Biochemical parameters

This study found that silver nanoparticle (AgNP) exposure significantly influenced blood glucose, triglyceride, cholesterol, and liver enzyme levels in *Oreochromis niloticus* (Table 3).

Blood glucose levels showed no significant changes in the control group or in fish exposed to lower AgNP concentrations (1, 3, and 6 μ g/L). However, a marked decrease was observed at the highest concentration of 12 μ g/L.

In contrast, triglyceride levels exhibited a significant, dose-dependent increase, beginning at 1 μ g/L and peaking at 267.1 mg/dL at 12 μ g/L.

Cholesterol levels demonstrated a reverse trend, with a significant decrease recorded at higher AgNP concentrations. Notable reductions were observed from 6 μ g/L, reaching the lowest value of 212.0 mg/dL at 12 μ g/L.

Regarding liver enzyme activity, alanine aminotransferase (ALT) levels showed a significant elevation starting from 3 μ g/L and continued to rise with increasing AgNP concentration. Meanwhile, aspartate aminotransferase (AST) levels increased significantly from the lowest tested concentration (1 μ g/L), plateaued after 3 μ g/L, and remained consistently high through 12 μ g/L.

 $^{^{}a}P < 0.01$ vs. control group

 $^{{}^{}b}P < 0.01 \text{ vs. } 1 \text{ µg/L of AgNPs treated group.}$

 $^{^{}c}P < 0.01 \text{ vs. } 3 \mu\text{g/L of AgNPs treated group.}$

 $^{^{}d}P < 0.01$ vs. 6 µg/L of AgNPs treated group.

 $^{^{}e}P < 0.01 \text{ vs. } 12 \text{ }\mu\text{g/L} \text{ of AgNPs treated group.}$

Table 3. Biochemical parameters of the blood plasma (mean \pm SE; n = 5) of the Nile tilapia (*Oreochromis niloticus*) exposed to AgNPs for 2 weeks

PARAMETER	CONTROL	1μG/L AGNPS	3 µG/L AGNPS	6 μG/L AGNPS	12 μG/L AGNPS
GLUCOSE (MG/DL)	34.0±15	33.4±2.9 ^{a,c,d,e}	34.1±5.0 a, b, d,	35.7±3.0 a, b, c, e	30.90±0.50 ^{a,b,c,d}
TRIGLYCERIDES (MG/DL)	188.1±0.60	221.4±0.80 ^a	252.9±0.70 a	262.1±0.60 a	267.1±0.50 ^a
CHOLESTEROL (MG/DL)	376.8±0.70	350.5±0.50 ^{d, e}	325.5±0.90 °	225.5±0.50 ^a ,	212.0±0.80 a, b, c
ALT	6.1±0.90	8.7 ± 0.9^{d}	9.6±0.5 a, d	11.3±0.7 a	14.0±0.11 a, b, e
AST	6.1±0.7	7.0±0.9 ^{a, c, d, e}	$8.7\pm0.8^{\mathrm{a,b}}$	8.7 ±0.7 ^{a, b}	8.7 ±0.7 ^{a, b}

ALT Alanine aminotransferase, AST Aspartate aminotransferase.

Data are expressed as mean \pm SEM. n = 8 for each experimental group

Gene expression

The data in Table (4) illustrate the impact of various concentrations of AgNPs on the expression of IL- $l\beta$ in the Nile tilapia (*Oreochromis niloticus*) as a fold-change following 30 days of exposure. The results indicate a progressive increase in IL- $l\beta$ expression with rising AgNPs concentrations, with the highest levels observed at the 12 μ g/L exposure, suggesting a heightened inflammatory response at this concentration.

Table 4. Effects of AgNPs on IL- 1β expression of Nile tilapia (*Oreochromis niloticus*)

PARAMETER	CONTROL	1 μG/L AGNP	3 μG/L AGNP	6 μG/L AGNP	12 μG/L AGNP
IL-1B	4.25±0.01	4.35±0.02 ^{a, c, d, e}	4.56±0.03 ^{a, b, c, d}	4.82±0.03 ^{a, b, c, e}	4.85±0.01 ^{a, b, d, e}
FOLD- CHANGE					

Data are expressed as mean \pm SEM. n = 8 for each experimental group

 $^{^{}a}P < 0.01$ vs. control group

 $^{{}^{}b}P < 0.01 \text{ vs. } 1 \text{ µg/L of AgNPs treated group.}$

 $^{^{}c}P < 0.01$ vs. 3 µg/L of AgNPs treated group.

 $^{^{}d}P < 0.01$ vs. 6 µg/L of AgNPs treated group.

 $^{^{}e}P < 0.01 \text{ vs. } 12 \text{ } \mu\text{g/L} \text{ of AgNPs treated group.}$

 $^{^{}a}P < 0.01$ vs. control group

 $^{{}^{}b}P < 0.01 \text{ vs. } 1 \text{ } \mu\text{g/L} \text{ of AgNPs treated group.}$

^cP < 0.01 vs. 3 µg/L of AgNPs treated group.

 $^{^{}d}P < 0.01$ vs. 6 µg/L of AgNPs treated group.

 $^{^{}e}P < 0.01$ vs. 12 µg/L of AgNPs treated group.

Histological examination

Fig. (1) presents the histological examination of liver tissues from *Oreochromis niloticus* exposed to varying concentrations of silver nanoparticles (AgNPs), revealing a dose-dependent pattern of hepatic alterations.

In the control group, liver architecture was well-preserved, with radiating cords of polygonal hepatocytes, centrally located nuclei, and clearly defined sinusoids.

At 1 and 3 µg/L of AgNP exposure, mild cytoplasmic vacuolation and scattered pyknotic nuclei were observed, indicating early signs of cellular stress.

At 6 μ g/L, moderate pathological changes became evident. These included marked cytoplasmic vacuolation, nuclear degeneration (such as pyknosis and karyorrhexis), dilated hepatic sinusoids, and mild leukocytic infiltration.

At the highest concentration of $12 \mu g/L$, the liver exhibited severe histopathological damage. This included extensive hepatocyte necrosis, intense cytoplasmic vacuolation, hemorrhagic regions, sinusoidal congestion, and dense leukocytic infiltration, resulting in near-total disruption of normal liver architecture.

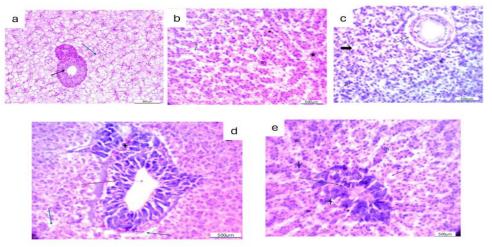


Fig. 1. Histological sections of tilapia liver from control and AgNPs exposed groups (H&E stain). (a) The control group displays normal hepatic architecture with polygonal hepatocytes, centrally located nuclei, and intact hepatic sinusoids. (b) Liver exposed to 1 μg of AgNPs shows mild cytoplasmic vacuolation (blue arrow) and early signs of nuclear pyknosis (strike). (c) Exposure to 3 μg induces hepatocellular degeneration (blue arrow), cytoplasmic vacuolization (black arrow), disorganization of hepatic cords, sinusoidal congestion (s), and prominent leukocytic infiltration. (d) At 6 μg exposure, hepatocytes exhibit moderate vacuolation, sinusoidal dilation, hemorrhage (strike), nuclear fragmentation, and inflammatory infiltration. (e) Exposure to 12 μg induces severe hepatocyte necrosis, widespread vacuolation (blue arrow), hemorrhage (strike), sinusoidal congestion (s) and complete disorganization of hepatic cords, indicating dosedependent hepatotoxicity. Scale bar = 500 μm

Histological examination of the tilapia intestine investigated distinct dose-dependent alterations in tissue architecture in response to AgNPs exposure, compared to the control group. In Fig. (2), sections of intestinal tissue from the control group exhibited normal histological architecture. The mucosal layer displayed well-formed, tall villi lined by columnar epithelial cells with abundant goblet cells interspersed. The lamina propria contained loose connective tissue with no signs of edema or inflammation. Intestinal sections exposed to 1 and 3 µg/l of AgNPs showed slight shortening and blunting of villi, occasional epithelial lifting, and reduced goblet cells. More pronounced histological changes were observed at 6 µg/l. The intestinal villi were markedly shortened and, in some areas, fused. The epithelial lining showed focal detachment, nuclear pyknosis, and cellular necrosis. Goblet cell numbers were further reduced. The lamina propria demonstrated moderate inflammatory infiltration and dilated blood vessels. Fish exposed to 12 µg/l exhibited severe histopathological alterations. These included extensive villus atrophy, epithelial sloughing, and loss of mucosal integrity. Goblet cells were nearly depleted, compromising mucosal defense. The lamina propria showed intense leukocyte infiltration, hemorrhage, and marked tissue disorganization. The submucosa was edematous.

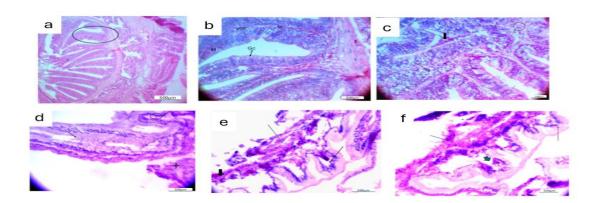


Fig. 2. Histological sections of tilapia intestine from control and AgNPs-exposed groups (H&E stain 20x & 40x). (a & b) The control group shows normal villi with intact epithelium and numerous goblet cells (Gc). (c & d) The intestines of tilapia exposed to $1 \mu g/l$ and $3 \mu g/l$ AgNPs exhibit mild villus shortening (green arrow), cytoplasmic vacuolation (thin arrow), reduced goblet cells (green arrow), submucosal fibrosis, and tissue disorganization (strike). (e) Exposure to $6 \mu g/L$ AgNPs results in villus fusion, epithelial necrosis (thick arrow), and lamina propria inflammation (black arrow). (f) Sections from intestines exposed to $12 \mu g/L$ AgNPs show severe epithelial sloughing (green arrow), villus atrophy (red arrow), submucosal fibrosis, and leukocytic infiltration (thick arrow)

DISCUSSION

The toxicity of silver nanoparticles (AgNPs) in aquatic environments is largely influenced by their bioavailability and capacity to infiltrate tissues. In this study, exposure to AgNPs in Nile tilapia (*Oreochromis niloticus*) led to several significant biochemical and physiological changes, particularly affecting lipid metabolism and liver function.

Increased triglyceride levels and decreased cholesterol levels were observed in a dose-dependent manner, highlighting disruptions in lipid homeostasis. These findings suggest an elevated energy demand in response to stress, consistent with previous research indicating that stress in fish leads to increased lipid mobilization (Lupatsch et al., 2010; Canli et al., 2018). Liver enzyme activities—ALT and AST—also significantly increased, indicating hepatocellular damage. These enzymes are recognized biomarkers for hepatotoxicity in fish, with elevated levels often linked to cellular injury and pathological conditions (Banaee et al., 2011; Fırat et al., 2011; Monfared & Soltani, 2013). Similar enzyme elevations have been reported following AgNP exposure in other species, including Labeo rohita and small mammals (Rajkumar et al., 2016; Vasanth & Kurian, 2017). These responses may reflect the liver's role as the primary organ for nanoparticle detoxification and accumulation (Ale et al., 2018b; Clark et al., 2019).

Interestingly, blood glucose levels remained relatively stable at lower AgNP concentrations, suggesting that glucose metabolism was not significantly disrupted under these conditions. However, studies have shown glucose increases at higher concentrations due to nanoparticle-induced stress (Farmen *et al.*, 2012; Valerio-García *et al.*, 2017). Elevated glucose is a well-established stress biomarker in fish and may result from disrupted endocrine regulation, including altered insulin secretion (Krishnasamy *et al.*, 2022).

The immune response was also markedly affected. A dose-dependent increase in IL- 1β expression was observed, indicating immune activation and inflammation. IL- 1β is a key pro-inflammatory cytokine associated with tissue damage and cellular stress. This is consistent with findings in other species where AgNP exposure increased inflammatory markers such as TNF- α and IL- 1β (Mansour *et al.*, 2021; Seo *et al.*, 2017; Bruneau *et al.*, 2016). Moreover, AgNPs have been linked to oxidative stress, DNA damage, and immunosuppression across a range of aquatic organisms, including mussels and zebrafish (Bruneau *et al.*, 2016; Bouallegui *et al.*, 2017).

Hematological parameters were significantly impacted by AgNP exposure. Increasing concentrations resulted in a marked reduction in RBC, WBC, and hemoglobin levels, indicating hematotoxicity. Similar outcomes have been reported in

Hypophthalmichthys molitrix and O. niloticus fed AgNP-contaminated diets (Shaluei et al., 2013; Rajkumar et al., 2016). These effects are likely linked to elevated reactive oxygen species (ROS) production, impaired hematopoiesis, and cell membrane damage. DNA damage in erythrocytes and WBC necrosis through caspase activation further support these findings (Orlowski et al., 2012; Bacchetta et al., 2017; Moustafa et al., 2025). Nonetheless, exposure to lower AgNP levels (e.g., 100 μg/L) has been shown to cause no significant hematological or histological alterations in some species, highlighting a threshold effect (Clark et al., 2019).

Histopathological examination provided further insight into tissue-level damage. Liver tissues showed dose-dependent alterations, beginning with mild cytoplasmic vacuolation at low concentrations (1–3 μ g/L), progressing to hepatocyte necrosis, hemorrhage, and architectural disruption at 12 μ g/L. These observations support the liver's critical role in nanoparticle detoxification and storage (**Khalil** *et al.*, **2019**; **Mansour** *et al.*, **2021**).

Similarly, intestinal tissues displayed a progressive pattern of damage. Mild effects such as villus blunting and epithelial lifting were seen at low concentrations, while high concentrations (12 µg/L) led to villus atrophy, goblet cell depletion, leukocytic infiltration, and submucosal edema, indicating severe mucosal disruption. Since the intestine plays a key role in nutrient and particle absorption, it is likely that AgNPs enter the bloodstream via intestinal uptake before reaching the liver (Isani et al., 2013). Prior studies have also confirmed the intestine as the second most affected organ following liver exposure (Zhao & Wang, 2011; Asghari et al., 2012; Moustafa et al., 2025). Similar histological damage to liver and pancreatic tissues has been reported in fish exposed to higher AgNP doses (1–2.5 mg/L) (Chakraborty et al., 2023).

CONCLUSION

This study demonstrates that silver nanoparticles (AgNPs) induce a range of dose-dependent hematological, biochemical, immunological, and histopathological alterations in *Oreochromis niloticus*. While low concentrations ($\leq 6~\mu g/L$) may be relatively safe for short-term exposure, higher levels ($12~\mu g/L$) are clearly associated with systemic toxicity, inflammation, and organ damage.

These findings highlight the need for caution in the application of AgNPs, especially in aquatic systems. Given their widespread industrial use and potential for environmental release, further research is essential to establish safe exposure thresholds and understand long-term ecological impacts. The Nile tilapia, due to their environmental sensitivity and economic importance, serve as an effective model for assessing nanoparticle toxicity and aquatic health.

REFERENCES

- Ale, A.; Bacchetta, C.; Rossi, A. S.; Galdopórpora, J.; Desimone, M. F.; de la Torre, F. R. and Cazenave, J. (2018a). Nanosilver toxicity in gills of a neotropical fish: metal accumulation, oxidative stress, histopathology and other physiological effects. *Ecotoxicology and Environmental Safety*, 148, 976-984.
- Ale, A.; Rossi, A. S.; Bacchetta, C.; Gervasio, S.; de la Torre, F. R. and Cazenave, J. (2018b). Integrative assessment of silver nanoparticles toxicity in *Prochilodus lineatus* fish. *Ecological Indicators*, 93, 1190-1198.
- Asghari, S.; Johari, S. A.; Lee, J. H.; Kim, Y. S.; Jeon, Y. B.; Choi, H. J. and Yu, I. J. (2012). Toxicity of various silver nanoparticles compared to silver ions in *Daphnia magna. Journal of nanobiotechnology*, 10, 1-11.
- **Asharani, P. V.; Wu, Y. L.; Gong, Z. and Valiyaveettil, S**. (2008). Toxicity of silver nanoparticles in zebrafish models. *Nanotechnology*, 19(25), 255102.
- Bacchetta, C.; Ale, A.; Simoniello, M. F.; Gervasio, S.; Davico, C.; Rossi, A. S. and Cazenave, J. (2017). Genotoxicity and oxidative stress in fish after a short-term exposure to silver nanoparticles. *Ecological indicators*, 76, 230-239.
- Banaee, M.; Sureda, A.; Mirvaghefi, A. R. and Ahmadi, K. (2011). Effects of diazinon on biochemical parameters of blood in rainbow trout (*Oncorhynchus mykiss*). *Pesticide biochemistry and physiology*, 99(1), 1-6.
- Blaise, C.; Gagné, F.; Ferard, J. F. and Eullaffroy, P. (2008). Ecotoxicity of selected nano-materials to aquatic organisms. *Environmental Toxicology: An International Journal*, 23(5), 591-598.
- **Bouallegui, Y.; Ben Younes, R.; Bellamine, H. and Oueslati, R.** (2017). Histopathology and analyses of inflammation intensity in the gills of mussels exposed to silver nanoparticles: role of nanoparticle size, exposure time, and uptake pathways. *Toxicology mechanisms and methods*, 27(8), 582-591.
- Bruneau, A.; Turcotte, P.; Pilote, M.; Gagné, F. and Gagnon, C. J. A. T. (2016). Fate of silver nanoparticles in wastewater and immunotoxic effects on rainbow trout. *Aquatic Toxicology*, 174, 70-81.
- **Burgos-Aceves, M. A.; Lionetti, L. and Faggio, C.** (2019). Multidisciplinary haematology as prognostic device in environmental and xenobiotic stress-induced response in fish. *Science of the total environment*, 670, 1170-1183.
- Cáceres-Vélez, P. R.; Fascineli, M. L.; Rojas, E.; Meyer, T.; Venus, T.; Grisolia, C. K. and Azevedo, R. B. (2019). Impact of humic acid on the persistence, biological fate and toxicity of silver nanoparticles: A study in adult zebrafish. Environmental Nanotechnology, Monitoring & Management, 12, 100234.
- Canli, T.; Yu, L.; Yu, X.; Zhao, H.; Fleischman, D.; Wilson, R. S. and Bennett, D. A. (2018). Loneliness 5 years ante-mortem is associated with disease-related

- differential gene expression in postmortem dorsolateral prefrontal cortex. *Translational psychiatry*, 8(1), 2.
- Chakraborty, P.; Krishnani, K. K.; Mulchandani, A.; Sarkar, D. J.; Das, B. K.; Paniprasad, K. and Pal, P. (2023). Toxicity assessment of poultry-waste biosynthesized nanosilver in *Anabas testudineus* (Bloch, 1792) for responsible and sustainable aquaculture development-A multi-biomarker approach. *Environmental Research*, 235, 116648.
- Choi, J. E.; Kim, S., Ahn, J. H.; Youn, P.; Kang, J. S.; Park, K. and Ryu, D. Y. (2010). Induction of oxidative stress and apoptosis by silver nanoparticles in the liver of adult zebrafish. *Aquatic Toxicology*, 100(2), 151-159.
- Clark, N.J.; Boyle, D.; Eynon, B.P. and Handy, R.D. (2019). Dietary exposure to silver nitrate compared to two forms of silver nanoparticles in rainbow trout: bioaccumulation potential with minimal physiological effects. Environ. Sci. Nano 6, 1393–1405.
- **Drake, P. L. and Hazelwood, K. J.** (2005). Exposure-related health effects of silver and silver compounds: a review. *Annals of Occupational Hygiene*, 49(7), 575-585.
- **El-Sayed Ali, T.; Abdel-Aziz, S. H.; El-Sayed, A. F. M. and Zeid, S.** (2014). Structural and functional effects of early exposure to 4-nonylphenol on gonadal development of Nile tilapia (*Oreochromis niloticus*): b-histological alterations in testes. *Fish physiology and biochemistry*, 40, 1495-1507.
- **El-Sayed, F. M. and Fitzsimmons, K.** (2023). From Africa to the world—the journey of Nile tilapia. Rev Aquac 15: 6–21.
- **FAO.** (2024). The state of world fisheries and aquaculture 2024: blue transformation in action.
- Farmen, E.; Mikkelsen, H. N.; Evensen, Ø.; Einset, J.; Heier, L. S.; Rosseland, B. O. and Oughton, D. H. (2012). Acute and sub-lethal effects in juvenile Atlantic salmon exposed to low μg/L concentrations of Ag nanoparticles. *Aquatic Toxicology*, 108, 78-84.
- Firat, Ö.; Cogun, H. Y.; Yüzereroğlu, T. A.; Gök, G.; Firat, Ö.; Kargin, F. and Kötemen, Y. (2011). A comparative study on the effects of a pesticide (cypermethrin) and two metals (copper, lead) to serum biochemistry of Nile tilapia, *Oreochromis niloticus*. Fish Physiology and Biochemistry, 37, 657-666.
- Gobi, N.; Vaseeharan, B.; Rekha, R.; Vijayakumar, S. and Faggio, C. (2018). Bioaccumulation, cytotoxicity and oxidative stress of the acute exposure selenium in *Oreochromis mossambicus*. *Ecotoxicology and environmental safety*, 162, 147-159.
- **Gottschalk, F.; Sun, T. and Nowack, B.** (2013). Environmental concentrations of engineered nanomaterials: review of modeling and analytical studies. *Environmental pollution*, 181, 287-300.

- Haney, D. C.; Hursh, D. A.; Mix, M. C. and Winton, J. R. (1992). Physiological and hematological changes in chum salmon artificially infected with erythrocytic necrosis virus. *Journal of Aquatic Animal Health*, 4(1), 48-57.
- **Ibrahim, A. T. A.** (2020). Toxicological impact of green synthesized silver nanoparticles and protective role of different selenium type on *Oreochromis niloticus*: hematological and biochemical response. *Journal of Trace Elements in Medicine and Biology*, 61, 126507.
- **Imani, R.; Emami, S. H. and Faghihi, S.** (2015). Synthesis and characterization of an octaarginine functionalized graphene oxide nano-carrier for gene delivery applications. *Physical Chemistry Chemical Physics*, 17(9), 6328-6339.
- **Inshakova, E. and Inshakov, O.** (2017). World market for nanomaterials: Structure and trends. In *MATEC web of conferences* (Vol. 129, p. 02013). EDP Sciences.
- Isani, G.; Falcioni, M. L.; Barucca, G.; Sekar, D.; Andreani, G.; Carpenè, E. and Falcioni, G. (2013). Comparative toxicity of CuO nanoparticles and CuSO4 in rainbow trout. *Ecotoxicology and environmental safety*, 97, 40-46.
- **Ji, J. H.; Jung, J. H.; Kim, S. S.; Yoon, J. U.; Park, J. D.; Choi, B. S. and Yu, I. J.** (2007). Twenty-eight-day inhalation toxicity study of silver nanoparticles in Sprague-Dawley rats. *Inhalation toxicology*, *19*(10), 857-871.
- **Johari, S. A.** (2014). Toxicity Effect of Colloidal Silver Nanoparticles on Fertilization Capacity and Reproduction Success of Rainbow Trout (*Oncorhynchus mykiss*). J Nanomed Res 1 (1): 00001.
- Kakavand, F.; Hedayati, A.; Nodeh, A. J.; Maddah, S. and Shadegan, M. R. (2020). Toxicological Effects of Silver Nanoparticles on Nile Tilapia (*Oreochromis niloticus*). Avicenna Journal of Environmental Health Engineering, 7(1), 1-7.
- **Khalil, H. S.; Mansour, A. T.; Goda, A. M. A. and Omar, E. A.** (2019). Effect of selenium yeast supplementation on growth performance, feed utilization, lipid profile, liver and intestine histological changes, and economic benefit in meagre, *Argyrosomus regius*, fingerlings. Aquaculture, 501, 135-143.
- Klaine, S. J.; Alvarez, P. J.; Batley, G. E.; Fernandes, T. F.; Handy, R. D.; Lyon, D. Y. and Lead, J. R. (2008). Nanomaterials in the environment: behavior, fate, bioavailability, and effects. *Environmental Toxicology and Chemistry: An International Journal*, 27(9), 1825-1851.
- Krishnasamy Sekar, R.; Arunachalam, R.; Anbazhagan, M.; Palaniyappan, S.; Veeran, S.; Sridhar, A. and Ramasamy, T. (2023). Accumulation, chronicity, and induction of oxidative stress regulating genes through Allium cepa L. functionalized silver nanoparticles in freshwater common carp (*Cyprinus carpio*). *Biological Trace Element Research*, 201(2), 904-925.
- Kumar, J.; Priyadharshini, M.; Madhavi, M.; Begum, S. S.; Ali, A. J.; Musthafa, M. S. and Faggio, C. (2022). Impact of Hygrophila auriculata supplementary diets on the growth, survival, biochemical and haematological parameters in fingerlings

- of freshwater fish Cirrhinus mrigala (Hamilton, 1822). *Comparative Biochemistry and Physiology Part A: Molecular & Integrative Physiology*, 263, 111097.
- **Lupatsch, I.; Santos, G. A.; Schrama, J. W. and Verreth, J. A. J.** (2010). Effect of stocking density and feeding level on energy expenditure and stress responsiveness in European sea bass *Dicentrarchus labrax*. *Aquaculture*, 298(3-4), 245-250.
- Ma, Y.; Song, L.; Lei, Y.; Jia, P.; Lu, C.; Wu, J. and Pei, D. S. (2018). Sex dependent effects of silver nanoparticles on the zebrafish gut microbiota. *Environmental Science: Nano*, 5(3), 740-751.
- **Mahmoudabadi, S.; Farahpour, M. R. and Jafarirad, S**. (2021). Effectiveness of green synthesis of silver/kaolinite nanocomposite using Quercus infectoria galls aqueous extract and its chitosan-capped derivative on the healing of infected wound. *IEEE Transactions on NanoBioscience*, 20(4), 530-542.
- Mansour, A.T.; Fayed, W.M.; Elkhayat, B.K.; Omar, E.A.; Zaki, M.A.; Nour, A.A.M. and Morshedy, S.A., 2021. Extract dietary supplementation affects growth performance, hematological and physiological status of European seabass. Ann. Anim. Sci. 21 (3), 1043–1060.
- Mansour, W. A.; Abdelsalam, N. R.; Tanekhy, M., Khaled, A. A. and Mansour, A. T. (2021). Toxicity, inflammatory and antioxidant genes expression, and physiological changes of green synthesis silver nanoparticles on Nile tilapia (*Oreochromis niloticus*) fingerlings. *Comparative Biochemistry and Physiology Part C: Toxicology & Pharmacology*, 247, 109068.
- Mohsenpour, R.; Mousavi-Sabet, H.; Hedayati, A.; Rezaei, A.; Yalsuyi, A. M. and Faggio, C. (2020). *In vitro* effects of silver nanoparticles on gills morphology of female Guppy (*Poecilia reticulate*) after a short-term exposure. *Microscopy research and technique*, 83(12), 1552-1557.
- **Monfared, A. L.; Soltani, S. and Louei, A.** (2013). Effects of silver nanoparticles administration on the liver of rainbow trout (Oncorhynchus mykiss): histological and biochemical studies. *European Journal of Experimental Biology*, *3*(2), 285-289.
- Moustafa, B. H.; Massoud, R. G. and Malash, M. A. (2025). The Nile Tilapia (*Oreochromis niloticus*) Physiological and Histopathological Biomarkers of some Fungal Infections and Proposed Silver Nanoparticles Remediation. *Egypt. J. Aquat. Biol. Fish*, 29(3): 1571 1591.
- **Ninan, N.; Goswami, N. And Vasilev, K.** (2020). The impact of engineered silver nanomaterials on the immune system. *Nanomaterials*, 10(5), 967.
- Orlowski, G.; Kamiński, P.; Kasprzykowski, Z. and Zawada, Z. (2012). Metal interactions within and between tissues of nestling rooks Corvus frugilegus. *Biologia*, 67, 1211-1219.

- **Rajkumar, K. S.; Kanipandian, N. and Thirumurugan, R. J. A. N.** (2016). Toxicity assessment on haemotology, biochemical and histopathological alterations of silver nanoparticles-exposed freshwater fish *Labeo rohita*. *Applied Nanoscience*, 6, 19-29.
- Ramírez-García, A.; Ramírez-Herrejón, J. P.; Medina-Nava, M.; Hernández-Morales, R. and Domínguez-Domínguez, O. (2018). Reproductive biology of the invasive species Pseudoxiphophorus bimaculatus and *Poecilia sphenops* in the Teuchitlán River, México. Journal of Applied Ichthyology, 34(1), 81-90.
- Rashidian, G.; Lazado, C. C.; Mahboub, H. H.; Mohammadi-Aloucheh, R.; Prokić, M. D.; Nada, H. S. and Faggio, C. (2021). Chemically and green synthesized ZnO nanoparticles alter key immunological molecules in common carp (*Cyprinus carpio*) skin mucus. *International Journal of Molecular Sciences*, 22(6), 3270.
- Savorelli, F.; Manfra, L.; Croppo, M.; Tornambè, A.; Palazzi, D.; Canepa, S. and Faggio, C. (2017). Fitness evaluation of *Ruditapes philippinarum* exposed to Ni. *Biological Trace Element Research*, 177, 384-393.
- Schaperclaus, W.; Kulow, H. and Schreckenbach, K. (1991). Hematological and serological technique. Fish disease, 1(56), 71-108.
- Seo, S. B.; Dananjaya, S. H. S.; Nikapitiya, C.; Park, B. K.; Gooneratne, R.; Kim, T. Y. and De Zoysa, M. (2017). Silver nanoparticles enhance wound healing in zebrafish (Danio rerio). Fish & shellfish immunology, 68, 536-545.
- **Shabrangharehdasht, M.; Mirvaghefi, A. and Farahmand, H.** (2020). Effects of nanosilver on hematologic, histologic and molecular parameters of rainbow trout (*Oncorhynchus mykiss*). *Aquatic Toxicology*, 225, 105549.
- **Shaluei, F.; Hedayati, A.; Jahanbakhshi, A.; Kolangi, H. and Fotovat, M.** (2013). Effect of subacute exposure to silver nanoparticle on some hematological and plasma biochemical indices in silver carp (*Hypophthalmichthys molitrix*). *Human & experimental toxicology*, 32(12), 1270-1277.
- Valerio-García, R. C.; Carbajal-Hernández, A. L.; Martínez-Ruíz, E. B.; Jarquín-Díaz, V. H.; Haro-Pérez, C. and Martínez-Jerónimo, F. (2017). Exposure to silver nanoparticles produces oxidative stress and affects macromolecular and metabolic biomarkers in the goodeid fish *Chapalichthys pardalis*. Science of the Total Environment, 583, 308-318.
- Vali, S.; Mohammadi, G.; Tavabe, K. R.; Moghadas, F. and Naserabad, S. S. (2020). The effects of silver nanoparticles (AgNPs) sublethal concentrations on common carp (*Cyprinus carpio*): Bioaccumulation, hematology, serum biochemistry and immunology, antioxidant enzymes, and skin mucosal responses. *Ecotoxicology and environmental safety*, 194, 110353.

- Waalewijn-Kool, P. L.; Klein, K.; Forniés, R. M. and van Gestel, C. A. (2014). Bioaccumulation and toxicity of silver nanoparticles and silver nitrate to the soil arthropod *Folsomia candida*. *Ecotoxicology*, 23, 1629-1637.
- **Yalsuyi, A. M. and Vajargah, M. F.** (2017). Acute toxicity of silver nanoparticles in Roach (*Rutilus rutilus*) and Goldfish (Carassius auratus). *Journal of Environmental Treatment Techniques*, 5(1), 1-4.
- Yalsuyi, A. M.; Hedayati, A.; Vajargah, M. F. and Mousavi-Sabet, H. (2017). Examining the toxicity of cadmium chloride in common carp (*Cyprinus carpio*) and goldfish (*Carassius auratus*). *Journal of Environmental Treatment Techniques*, 5(2), 83-86.
- Yan, L.; Zheng, Y. B.; Zhao, F.; Li, S.; Gao, X.; Xu, B. and Zhao, Y. (2012). Chemistry and physics of a single atomic layer: strategies and challenges for functionalization of graphene and graphene-based materials. *Chemical Society Reviews*, 41(1), 97-114
- **Zhao, C. M. and Wang, W. X.** (2011). Comparison of acute and chronic toxicity of silver nanoparticles and silver nitrate to *Daphnia magna*. *Environmental Toxicology and Chemistry*, 30(4), 885-892.