# Elements Accumulation and Nutritive Value of *Phragmites Australis* (Cav.) Trin. ex Steudel in Lake Burullus: A Ramsar site, Egypt

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#### ABSTRACT

The present study aims to assess the role of *Phragmites australis* in the accumulation of elements and nutritive value to test its suitability to use as a potential forage plant for cattle, goats and sheep in lake Burullus, Egypt. Different plant organs were collected from the sampling sites for estimating seven heavy metals (Mn, Pb, Co, Cu, Cd, Zn, and Fe) and six nutrients (Na, K, Ca, Mg, P and N) as well as the physical and chemical characteristics of water. In addition, seven organic constituents (total ash, carbohydrates, total lipids, crude fibers, crude protein, digestible crude protein and total digestible nutrients) were estimated, and three nutritive values (digestible energy, metabolized energy, net and gross energy) were calculated. The results revealed that an increase in heavy metals accumulation in the rhizome and decreased in the order of rhizome > stem > leaves. Moreover, positive linear relationships were found between these heavy metal concentrations in plant organs and those in water. Thus P. australis can serve as a good accumulator and bioindicator of heavy metals in the polluted water bodies. On the other hand, the nutrients are decreased in the order of stems > leaves > rhizomes. The leaves had the highest total ash, crude protein, digestible crude protein, digestible energy, metabolized energy and net energy, while the stems had the highest total carbohydrates and crude fibers. Therefore, the results revealed that the underground rhizomes had the ability to accumulate heavy metals and thereby used as a phytoremediator; while its aboveground parts had the highest nutrient and nutritive values, which consider the plant as good forage for animals.

Key words: Bioaccumulation, Common reed, Forage plant, Heavy metals, Phytoremediation.

## INTRODUCTION

The common reed, *Phragmites australis* (Cav.) Trin. ex Steudel, is a cosmopolitan angiosperm believed to be one of the most widely distributed and valuable species in the world (Holm *et al.*, 1977). In Egypt, it is known to be the major component of reed stands along the shores of Lake Burullus. Green plants of *P. australis* are considered very palatable and readily grazed by sheep, goats, cattle, wildlife and provide important shade, shelter and food for fishes and that common reed litter provides food for molluscs, other crustaceans and aquatic insects (Holm *et al.*, 1977; Frankenber, 1997; Shaltout and Al-Sodany, 2008).

Although its stands can maintain important eco-system functions, it is regarded as a harmful plant in other regions (Weis and Weis, 2003; and Bonanno and Giudice, 2010). Thus, different aspects of this species such as biology and ecology have been extensively studied during the last decades. However, until now its nutrient budget in the Mediterranean region receives little attention (Eid *et al.*, 2010b). Information on these aspects is important for deriving sound management recommendations for *P. australis* stands aiming at optimizing carbon sequestration and counteracting eutrophication. Phytoremediation means the use of plants to reduce, remove, degrade or immobilize environmental toxins (Raskin *et al.*, 1997; Salt *et al.*, 1998; Terry *et al.*, 2003).

The effectiveness of a phytoremediation system depends on the selection of appropriate plants for the particular environment. Knowledge about the accumulation properties of wetland plant species is useful in choosing appropriate plants for wetland phytoremediation systems (Duman et al., 2007). Metal bioaccumulation depends upon numerous biotic and а biotic factors, such as temperature, pH and dissolved ions in water (Lewander et al., 1996; Demirezen and Aksoy, 2004). Bioaccumulation of metals varies considerably among species growing in the same area, as well as within the same species during different seasons (Brekken and Steinnes, 2004). Some studies have reported the highest metal contents (Cd. Cu. Ni, Pb. Sn. Zn) during autumn and relatively low levels during spring (Brekken and Steinnes, 2004), whereas others have indicated the highest levels during spring and the lowest during winter (Wilkins, 1978; Martin and Couphtrey, 1982).

*P* australis has been widely investigated as a bioabsorbent of nutrients and heavy metals (Ruiz and Velasco, 2010). It functions as a filter for reducing pollution sources (Brix and Schierup, 1989; Kiedrzynska *et al.*, 2008). It is one of the emergent plants most commonly used in constructed wetlands for the enhancement of water quality in water treatment systems (Gómez *et al.*, 2000; Borin *et al.*, 2001; Meuleman *et al.*, 2002; Vyamazal, 2002) due to its high growth rate and



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great capacity for nutrient accumulation in its organs (Asaeda *et al.*, 2002; Baldatoni *et al.*, 2003; Kiedrzynska *et al.*, 2008).

The capacity of natural wetlands to filter waste water and nutrients from arable land also has been demonstrated (Newman and Pietro, 2001; Cirujano *et al.*, 2005; Álvarez-Rogel *et al.*, 2006). Besides, reed beds have other ecological and conservation value and important wetland functions, such as a major breeding habitat for passerines (Poulin *et al.*, 2002) and contributing margin sediment stabilization to aquatic biodiversity (Shaltout and Khalil, 2005). However, reeds tends to form monospecfic and dominant stands and may alter ecological functions by their excessive expansion, adopting an invasive behaviour (Lelong *et al.*, 2007), especially in coastal marshes, where there are increase of inputs of freshwater and nutrient loads and the absence of other species (Minchinton and Bertness, 2003).

In Egypt, the serious problem of feed shortage, especially green summer fodder suppresses the improvement of animal production. Therefore, dependence on improving local food and food resources for both animals and humans is necessary for a sound policy. The animal feeding system is depending on the cultivation of Egyptian clover (*Trifolium alexandrinum L.*). It produces around 4.8 million ton of starch year<sup>-1</sup> and could cover the requirements of animals with surplus of 0.9 million ton of starch. However, in summer period, there will be at least a deficiency of 1.5 million ton of starch and most animals are in fact in a starving condition receiving less than the normal requirements (Gabra *et al.*, 1987). This calls for studying other non-conventional sources of feed.

Little quantitative information is available about the importance of *P. australis* in the nutrient budget of reed marshes and the possible role of this species in the phytoremediation of the Mediterranean wetlands. It is therefore important to evaluate the spatial and organ variation in plant accumulation in wetland system in order to assess the potential capacity of nutrient and metal removal by plant uptake. The present study aims to evaluate the role of *P. australis* in the accumulation of seven heavy metals (Mn, Pb, Co, Cu, Cd, Zn and Fe) and six nutrients elements (Na, K, Ca, Mg, P and N) in Lake Burullus, a Ramsar site along the Deltaic Medit-erranean coast of Egypt.

It also aims to estimate the variation in seven organic constituents (total ash, carbohydrates, total lipids, crude fibres, crude protein, digestible crude protein and total digestible nutrients) and their three nutritive values (digestible energy, metabolized energy, net and gross energy) were calculated to evaluate their forage quality in relation to the space and plant organ.

## MATERIALS AND METHODS

### Study area

Lake Burullus is one of the Egyptian northern lakes natural outlets connecting with the Mediterranean Sea.

This shallow lake has an oblong shape extending for a distance of 47 km along NE-SW axis. Its main basin is classified into three sections: eastern, middle and western. The western section has less width not exceeding 5 km, and increases in the middle section to reach a maximum of 14 km (Fig. 1).

The depth of this lake varies between 20 cm close to the shore of its eastern section and 200 cm at the middle section and near the sea outlet. A marine sand bar separates the Mediterranean coast from the lake shore, with a depth varying between few hundred meters near the sea outlet and a maximum of 6 km in the west. Some 25 islets of different sizes are distributed within the lake where they form physical isolations between the 3 sections of the lake. The heavy growth of reed and sedge plants (e.g. *P. australis* and *Typha domingensis*) facilitates the merging of the nearby islets (Shaltout and Khalil, 2005).



Figure (1): Map of Lake Burullus indicating the location of the eight sampling sites (\*).

The Mediterranean Deltaic coast, in which Lake Burullus occupies, belongs to the arid region where the climatic conditions are warm in summer (20 - 30°C) and mild in winter (10 - 20°C) with an aridity index ranging between 0.20 and 0.03 (UNESCO, 1977). Generally, the days are sunny and dry. The maximum and minimum mean annual temperatures are 26.5 °C and 14.4 °C. The mean relative humidity is about 69%, the mean evaporation is 4.6 mm day<sup>-1</sup> and the mean rainfall is about 175 mm year <sup>-1</sup> (Anonymous, 1980).

#### Plant analysis

The sampling process was carried out in the eastern, middle and western section (E-W) of Lake Burullus. In each section, three sites were selected to represent the north, middle and south (N-S) of the lake, except that of western section where only two sites were selected: one in the north and the other in the south (Fig. 1).

These sites have pure or nearly pure *Phragmites* stands

with other aquatic species. In each site, three randomly distributed quadrates (each of  $0.5 \ge 0.5$  m size) were sampled with in *P. australis* population. For heavy metal and nutrient contents, samples of the different plant parts (leaves, stems and rhizomes) were collected from the sampling quadrates (12 replicates for each analysis), washed gently once with tap water, and three times with deionised water. Afterwards, the materials were ovendried (60°C) to constant weight, and powdered in a metal-free plastic mill.

Total ash (TA) percentage was estimated by ignition at  $550^{\circ}$ C for 2 hrs. Heavy metal and nutrients were extracted from 0.5-1 g samples (leaves, stems and rhizomes) using mixed-acid digestion method. Na and K were analyzed using the flame photometer; Ca, Mg, Mn, Pb, Co, Zn, Cu, Fe and Cd by atomic absorption and P and N by spectrophotometer. All these procedures were according to Allen *et al.* (1989). Total lipids (TL) and crude fibre (CF) were determined by ether extracting with soxhlet extraction method (Allen *et al.*, 1989). Crude protein (CP) was calculated following this equation (Öelberg, 1965): CP = IN × 6.25, where IN is insoluble nitrogen.

Digestible crude protein (DCP) was calculated according to the equation of Demarquilly and Weiss (1970): DCP (in % DM) = 0.929 CP (in % DM) - 3.52, where DM is the dry matter. Carbohydrates (Car) or (NFE) were calculated according to Le Houérou (1980): Car (in % DM) = 100 - (CP + CF + TL + MINS), where MINS is the total minerals.

Total digestible nutrients (TDN) were estimated according to the equation applied by (Naga and El-Shazly, 1971): TDN (in % DM) = 0.62 (100 + 1.25 TL) - PK, where TL is the percentage of total lipids or ether extract EE, P is the percentage of crude protein, and K is the coefficient that depends on the protein and fibre contents (0.7). Digestible energy (DE) was estimated following this equation (NRC, 1984b): DE (Mcal kg<sup>-1</sup>) = 0.0504 CP (%) + 0.077 TL (%) + 0.02 CF (%) + 0.000377 (NFE)<sup>2</sup> (%) + 0.011 NFE (%) - 0.152. Metabolized energy (ME) equals 0.82 DE, while net energy (NE) equals 0.5 ME (Garrett, 1980). Gross energy (GE) was calculated following this equation (Garrett, 1980): GE (Kcal kg<sup>-1</sup>) = (5.72 CP +9.5 EE + 4.79 CF +4.03 NFE)/10.

#### Water analysis

In each site, water depth and water transparency were measured using Seechi disc with a diameter of 30 cm. The disc was lowered slowly from the shaded side of the boat till it disappears for transparency, and lowered again till setting up the bottom of the depth. Water

salinity (as electric conductivity) and pH were estimated, in situ, using conductivity and pH meters.

Three water samples were collected from each site, in polyethylene bottles. Dissolved oxygen was estim-ated using Winkler's or iodometric method, biochemical oxygen demand (BOD) using 5 day BOD test, and chemical oxygen demand (COD) using titrimetric method. Alkalinity was determined using the titration against HCl and chlorosity using AgNO<sub>3</sub>. Dissolved nitrite, nitrate and phosphate were determined using spectrophotometric methods; while reactive silicate using heteropoly blue method. The heavy metals Cu, Fe, Zn, Cd and Pb were estimated in acidified water samples using atomic absorption method. All these methods are outlined in detail in Greenberg *et al.* (1992).

#### Data analysis

Elements and nutritive data for *P. australis* organs were subjected to a two-way analysis of variance (ANOVA-2) to test the differences between plant organs over spatial variations. The significant of variation in water characteristics over spatial variation was assessed using one-way analysis of variance (ANOVA-1). Simple linear correlation coefficients were calculated to assess the relationship between the population and water characteristics using SPSS software (SPSS, 2006).

#### RESULTS

Rhizomes exhibit the highest concentrations of heavy metals Mn, Pb, Co, Cu, Cd, Zn and Fe (102.3 $\pm$ 12.2, 95.5 $\pm$ 9.7, 11.2 $\pm$ 3.7, 22.7 $\pm$ 16.5, 23.4 $\pm$ 18.5, 306.5 $\pm$ 33.6 and 2134.4 $\pm$ 1338.4 µg g<sup>-1</sup> dry wt, respe-ctively) (Table1).

For all organs, the heavy metals (Mn, Pb and Co) differ significantly between the different sections of the lake. On the other hand, stem exhibits the highest concentrations of nutrients Na, K, Ca and Mg (8.3±1.6, 13.7±2.5, 8.5±2.1 and  $6.6\pm1.4$  mg g<sup>-1</sup> dry wt, respectively); while leaves exhibit the highest of total nitrogen (2.33± 0.49 mg g<sup>-1</sup> dry wt.) (Table 2). For all organs, the nutrients (Na, K, Ca, Mg and P) differ significantly between the different sections of the lake (E-W and N-S) (P < 0.001); while for the rhizomes Na (F = 5.1, P < 0.05), K (F = 7.4, P < 0.01), Ca (F = 10.6, P < 0.01) and P (F = 20.8, P < 0.001) differ significantly between different sites.

The total nitrogen differs significantly between the sections (F = 29.0, P < 0.001) sites (F = 11.5, P < 0.001) and the interaction (F = 8.6, P < 0.01) between them for leaves. In general, the heavy metals of *P. australis* organs have the following order: Fe > Zn > Mn > Pb > Cu > Cd > Co while their nutrients have the following order: K > Ca > Na > Mg > N > P (Fig. 2 and 3).

Regarding the water characters in the E-W section of the lake, the west site had the highest values of water transparency (42.2 cm), depth (145.6 cm), COD (5.7 mg  $\Gamma^{-1}$ ), BOD (4.4 mg  $\Gamma^{-1}$ ), and NO<sub>3</sub>, NO<sub>2</sub>, SiO<sub>2</sub>, Cu, Fe and Pb (3.2, 1.8, 37.1, 5.9, 2.0 and 2.4 µg  $\Gamma^{-1}$ , respectively), but the lowest of EC (2.0 mS cm<sup>-1</sup>), pH (8.5), alkalinity (211.8 mg  $\Gamma^{-1}$ ), Zn (2.8 µg  $\Gamma^{-1}$ ) and Cl (0.7 g  $\Gamma^{-1}$ ) (Table 3). On the other hand, in the N-S section, the south site had the highest values of alkalinity, COD and BOD (285.3,5.7 and 4.4 mg  $\Gamma^{-1}$ , respectively), and PO<sub>4</sub>, NO<sub>3</sub>, NO<sub>2</sub>, Cu, Fe, Cd and Pb (1.8, 3.1, 1.6, 5.2, 1.6, 2.4 and 1.9 µg  $\Gamma^{-1}$ , respectively); but the lowest of water transparency (28.1 cm), EC (4.9 mS cm<sup>-1</sup>) and SiO<sub>2</sub> (19.1 µg  $\Gamma^{-1}$ ); whereas, the north site in this section N-S had the

vic versa of these variables. The simple linear correlations between the heavy metals in the water and the plant organs indicate significant positive correlation of lead in water and all organs (r = 0.82 - 0.98, P < 0.05 -< 0.001) and Fe in water, stem (r = 0.89, P < 0.01) and rhizome (r = 0.93, P< 0.01). Also, the lead has significant positive correlation between the plant organs (r = 0.83 - 0.96, P < 0.05 - < 0.01), and Zinc has significant positive correlation (r = 0.92, P < 0.01) in leaves and stem (Table 4). The leaves attain the highest percentages of total ash (14.15±1.43%), crude protein (14.56±3.10%) and digestible crude protein  $(10.04\pm2.91\%)$ ; stems attain the highest concentrations of total carbohydrates (52.99±4.99%) and crude fibers (35.91±4.46%); and rhizomes attain the highest of total lipids (3.94±0.77%) (Table 5). The organic contents exhibited the following order: carbohydrates > crude fibers > crude protein > total ash> total lipids (Table 5 and Fig. 4).

For the leaves, carbohydrates, crude protein and digestible crude protein differ significantly between the sections and sites of the lake and the interaction between them (P < 0.001); while for the stem (P < 0.01) and rhizome (P < 0.001), the total ash differs significantly between the sites only. On the other hand, the total lipids, carbohydrates and crude fibers differ significantly between the sections and sites for stem (P < 0.01) and rhizome (P < 0.001).

Regarding the nutritive value, the rhizomes attain the highest percentages of total digestible nutrients and gross energy ( $49.70\pm0.53\%$  and  $42.71\pm0.70$  Mcal kg<sup>-1</sup>); while the leaves attain the highest concentrations of digestible energy, metabolized energy and net energy

The gross energy differ significantly between the sections and sites and the interaction between them for leaves (P < 0.001); while digestible energy, metabolized energy and net energy differ significantly between the sections and the interaction between sections and sites for stem and rhizome (Table 6).

#### DISCUSSION

Although, there are a little data on the use of littoral plants as heavy metals bioaccumulators over large areas of the wetlands environment, these littoral plants are commonly used in constructing wetlands for accumulating large amounts of heavy metals (Peverly *et al.*, 1995; Southichak *et al.*, 2006; Liu *et al.*, 2007; Vymazal *et al.*, 2009; Drzewiecka *et al.*, 2010 and Eid *et al.*, 2012).

In the present study, the heavy metal and nutrient concentrations were lower in the water than that in *P. australis* tissues; this indicating the excessive accumulation of the estimated elements in its tissues.

This suggestion is supported by the heavy metal accumulates in the rhizome more than in the stem and leaves. Many authors have similar results and stated that the underground parts of littoral plants such as *P. australis* (Vymazal *et al.*, 2007), *Typha angustifolia* L. and *Potamogeton pectinatus* L. (Demirezen and Aksoy, 2004), *Carex rostrata Stokes* (Stoltz and Greger, 2002),



**Figure (2):** Mean and standard deviation (vertical bars) of the heavy metals in the different organs of P. australis in Lake Burullus.



Figure (3): Mean and standard deviation (vertical bars) of the nutrienNB ts in the different organs of *P. australis* in Lake Burullus.



Figure (4): Mean and standard deviation (vertical bars) of the organic contents in the different organs of *P. australis* in Lake Burullus. TA: total ash, Car: total carbohydrates, TL: total lipids, CF: crude fibres, CP: crude protein, DCP: digestible crude protein and TDN: total digestible nutrients.

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	Section (Sc)	Site (St)	Mn	Pb	Со	Cu	Cd	Zn	Fe
		West	77.3±3.7	33.1±6.4	3.6±3.1	35.3±18.9	27.1±27.2	298.2±29.0	913.4±169.0
	E-W	Middle	$80.2\pm5.7$	$32.5 \pm 4.4$	$3.2 \pm 3.5$	29.4±10.9	15.7±17.4	$287.4\pm24.7$	791.0±97.2
		East	75.0±4.7	38.7±5.2	$4.6 \pm 2.2$	36.8±19.9	15.1±11.5	290.3±26.8	782.2±69.2
		South	$76.9\pm5.4$	34.0±5.6	$2.8 \pm 3.5$	35.7±14.8	$25.9 \pm 19.8$	294.6±33.9	837.6±167.1
Leaf	N-S	Middle	80.1±5.6	35.6±6.3	$3.8\pm2.4$	25.1±11.6	$11.5 \pm 14.9$	$278.3 \pm 17.1$	825.2±129.6
		North	76.3±4.7	$35.5 \pm 6.2$	$4.9\pm2.4$	$34.1 \pm 18.8$	$14.4{\pm}19.9$	$293.7{\pm}18.2$	807.3±69.7
			77.5±5.2	$35.0\pm5.9$	$3.8\pm2.9$	33.2±16.0	$18.8 \pm 19.5$	291.3±26.0	823.5±125.9
	Total	Sc	10.3***	9.6**	5.5**	0.7	2.3	0.3	3.5
	<b>F-Value</b>	St	5.1**	0.5	12.6***	0.5	2.7	0.4	0.6
		Sc*St	1.0	4.6*	11.3***	3.1	3.3*	0.5	2.0
		West	82.6±5.3	$64.4{\pm}17.2$	5.4±3.7	32.5±16.9	23.9±16.4	326.9±37.7	861.8±158.8
	E-W	Middle	93.8±12.9	$58.2 \pm 14.0$	$6.0\pm4.6$	18.7±6.9	13.5±13.9	$283.0\pm25.4$	791.8±114.4
		East	83.0±6.6	$77.8 \pm 4.8$	7.5±0.6	30.0±19.7	9.8±19.2	$289.4 \pm 20.1$	924.2±281.4
Stem		South	$84.9 \pm 6.0$	63.0±15.7	$6.0\pm3.4$	28.5±21.3	16.3±15.4	$305.6 \pm 30.9$	825.0±132.2
	N-S	Middle	93.3±15.3	$67.3 \pm 17.1$	$5.5 \pm 3.5$	$20.9 \pm 11.0$	$21.5 \pm 14.2$	$269.5 \pm 45.6$	697.8±70.5
		North	$87.3 \pm 8.1$	71.1±12.3	$6.9 \pm 3.4$	$27.4{\pm}16.3$	$14.4{\pm}16.4$	$294.7 \pm 33.8$	$876.2 \pm 183.0$
			87.9±10.0	67.1±14.8	6.2±3.4	26.7±17.5	16.4±15.3	294.8±35.9	824.3±157.3
	Total	Sc	$10.8^{***}$	50.3***	8.1**	2.5	1.3	7.6**	0.6
	F-Value	St	2.9	8.4***	10.1***	0.0	0.9	1.3	2.2
		Sc*St	1.7	5.1**	38.4***	2.2	0.7	2.3	0.5
		West	$98.3 \pm 8.7$	96.8±13.5	$9.5 \pm 7.4$	30.1±1.1	23.7±0.7	$325.2\pm6.9$	2373.4±489.5
	E-W	Middle	$106.0{\pm}14.4$	$89.8 \pm 9.0$	$10.6\pm2.4$	$20.8 \pm 16.0$	$29.3 \pm 15.6$	$297.3\pm26.7$	1125.8±328.3
		East	99.4±9.9	$101.7 \pm 4.1$	$12.7 \pm 2.9$	$23.4{\pm}18.6$	$21.2 \pm 19.9$	$312.5 \pm 40.6$	3103.2±1340.7
<b>Rhizome</b>		South	$106.9 \pm 13.6$	92.9±12.2	$10.4 \pm 4.7$	$26.7 \pm 14.2$	$37.4 \pm 12.6$	$302.9 \pm 25.0$	$2248.3 \pm 1365.5$
	N-S	Middle	103.7±13.3	95.0±7.7	$10.6\pm2.2$	$23.8 \pm 17.0$	$22.3\pm22.5$	$313.8 \pm 49.1$	$1890.4 \pm 1560.1$
		North	97.1±8.8	$98.1 \pm 8.8$	12.4±3.6	$18.7 \pm 18.4$	13.0±11.9	$303.4\pm26.5$	2238.5±1247.3
			102.3±12.2	95.5±9.7	11.2±3.7	22.7±16.5	23.4±18.5	306.5±33.6	2134.4±1338.4
	Total	Sc	2.4	21.8***	19.3***	0.3	3.3	0.8	16.8***
	F-value	St	3.3	2.8	11.9***	1.0	3.9	0.4	0.4
		Sc*St	1.1	2.5	30.7***	3.3	2.0	0.9	1.0

**Table (1):** Spatial variation in the mean concentration  $\pm$  standard deviation of the heavy metals ( $\mu$ g g<sup>-1</sup> dry wt) in*P. australis* in Lake Burullus. Sc: section, St: site,\*; P0.05, \*\*; P0.01, \*\*\*; P0.001

**Table (2)**: Spatial variation in the mean concentration ± standard deviation of the differentnutrients (mg g <sup>-1</sup> drywt) in *P. australis* in Lake Burullus. \*;P0.05, \*\*;P0.01, \*\*\*;P0.001.

	Section (Sc)	Site (St)	Na	K	Ca	Mg	Р	Ν
	E-W	West	5.6±0.6	9.8±1.2	6.0±1.7	4.4±0.7	0.3±0.1	2.19±0.55
		Middle	9.2±0.9	$14.7 \pm 1.1$	9.2±1.7	$6.6 \pm 1.0$	$0.2\pm0.1$	$2.54 \pm 0.37$
		East	$8.0\pm0.7$	$14.2 \pm 1.8$	$8.4{\pm}1.8$	$6.4\pm0.8$	$0.2\pm0.1$	2.21±0.53
<u>Leaf</u>	N-S	South	7.6±1.7	13.1±3.1	$7.9\pm2.5$	$5.9 \pm 1.4$	$0.2\pm0.1$	$2.34\pm0.54$
		Middle	8.7±1.3	$14.7 \pm 1.0$	8.9±1.5	6.5±0.9	$0.2\pm0.1$	$2.22\pm0.49$
		North	7.5±1.5	$12.6\pm2.3$	$7.8\pm2.0$	$5.6 \pm 1.1$	$0.2\pm0.1$	$2.40 \pm 0.47$
			7.8±1.6	13.3±2.5	8.1±2.1	5.9±1.2	$0.2 \pm 0.1$	2.33±0.49
	Total	Sc	106.1***	59.9***	63.8***	27.9***	19.5***	29.0***
	F-Value	St	0.5	1.1	0.3	0.6	5.1**	11.5***
		Sc*St	4.3*	5.7**	4.3*	2.7	7.7**	8.6**
	E-W	West	6.2±0.9	$10.6 \pm 1.2$	$6.5 \pm 1.6$	4.9±0.6	$0.2\pm0.1$	$0.85 \pm 0.17$
		Middle	$9.0{\pm}1.1$	$15.2 \pm 1.7$	$9.0{\pm}1.8$	$6.9 \pm 1.0$	$0.2\pm0.1$	$0.79 \pm 0.07$
		East	$7.8\pm0.9$	12.1±1.5	$8.2 \pm 1.1$	5.9±0.9	$0.2\pm0.1$	$0.78 \pm 0.09$
Stem	N-S	South	$8.0\pm2.0$	$13.5 \pm 2.7$	$8.2\pm2.5$	$6.5 \pm 1.9$	$0.2\pm0.1$	$0.90 \pm 0.20$
		Middle	$8.8\pm0.5$	15.4±1.9	9.1±2.0	7.3±0.9	$0.2\pm0.0$	0.94±0.34
		North	8.3±1.6	$12.8 \pm 2.2$	$8.4{\pm}1.8$	6.1±1.1	$0.2\pm0.1$	$0.76 \pm 0.08$
			8.3±1.6	13.7±2.5	8.5±2.1	6.6±1.4	$0.2 \pm 0.1$	$0.85 \pm 0.22$
	Total	Sc	136.6***	34.9***	72.6***	36.5***	4.8*	1.8
	<b>F-Value</b>	St	2.4	3.0	10.6**	1.9	0.6	2.7
		Sc*St	44.5***	5.3	29.4***	13.5***	3.2*	0.3
	E-W	West	$5.3\pm0.4$	$7.6\pm0.9$	$5.8 \pm 1.7$	$4.4\pm0.6$	$0.2\pm0.0$	$0.64 \pm 0.03$
		Middle	$7.6 \pm 1.0$	$12.7 \pm 1.1$	$7.7 \pm 1.8$	$5.6\pm0.6$	$0.1\pm0.0$	$0.63 \pm 0.04$
		East	7.1±0.9	$11.8 \pm 1.0$	7.0±1.3	$5.5\pm0.8$	$0.1\pm0.0$	$0.63 \pm 0.05$
<u>Rhizon</u>	N-S	South	7.1±1.3	$12.2\pm2.5$	$7.2 \pm 1.6$	$5.3 \pm 1.0$	$0.1\pm0.0$	$0.65 \pm 0.03$
		Middle	$6.9\pm0.5$	11.7±0.9	$7.0\pm1.7$	$5.7\pm0.5$	$0.1\pm0.0$	$0.62 \pm 0.05$
		North	7.1±1.5	$11.0\pm 2.3$	$7.2 \pm 1.9$	$5.2\pm0.8$	$0.2\pm0.0$	$0.62 \pm 0.04$
			7.1±1.2	$11.6 \pm 2.1$	7.2±1.7	5.4±0.8	$0.1 \pm 0.0$	$0.63 \pm 0.04$
	Total	Sc	34.9***	48.1***	72.6***	12.1***	8.8**	0.3
	F-value	St	5.1*	7.4**	10.6**	0.8	20.8***	0.9
		Sc*St	11.6***	0.1	29.4***	6.3**	8.6**	1.0

Table (3): Mean $\pm$ standard error (SE) of the water characters in Lake But	rulluss. DO: Di	ssolved Oxyge	n, COD:
Chemical Oxygen Demand, BOD: Biochemical Oxygen Demand, *;P	0.05, **;P	0.01, ***;P	0.001.

Element			E-W		F-value		N-S		Mean	F-value
		East	Middle	West	_	North	Middle	South	±SE	
Air ten	ър. (°С)	26.5	24.1	27.6	5.1**	26.1	25.4	26.2	$25.9\pm0.4$	0.3
Water te	mp. (°C)	26.9	26.5	26.9	0.1	26.4	27.0	26.8	$26.8\pm0.5$	0.1
Transpar	ency (cm)	29.9	36.2	42.2	7.4***	40.4	37.2	28.1	34.4±1.2	9.7***
Depth	( <b>cm</b> )	110.4	131.3	145.6	2.3	142.9	115.3	119.7	124.4±5.7	1.4
EC (m	S cm <sup>-1</sup> )	8.2	2.9	2.0	21.8***	6.0	4.9	4.9	5.2±0.6	0.3
<b>p</b> ]	H	9.1	8.8	8.5	10.4***	8.7	8.9	8.8	$8.8\pm0.1$	2.0
Alka	linity	285.5	292.3	211.8	14.1***	251.0	275.9	285.3	273±6.7	1.9
DO	(mg l <sup>-1</sup> )	7.9	7.8	8.4	0.2	8.7	7.7	7.9	$8.0{\pm}1.0$	1.1
COD	(mg l <sup>-1</sup> )	5.4	5.0	5.7	0.6	5.4	4.8	5.7	5.3±0.2	1.5
BOD	(mg l <sup>-1</sup> )	4.3	3.5	4.4	0.9	3.8	3.8	4.4	4.0±0.2	0.5
PO <sub>4</sub>	(µg l <sup>-1</sup> )	1.6	0.9	1.3	3.7*	0.7	1.2	1.8	1.3±0.1	9.6***
NO <sub>3</sub>	(µg l <sup>-1</sup> )	2.1	2.0	3.2	3.3*	1.3	2.1	3.1	2.3±0.1	11***
$NO_2$	(µg l <sup>-1</sup> )	1.0	1.0	1.8	4.4**	0.8	0.9	1.6	1.2±0.1	4.7**
SiO <sub>2</sub>	(µg l <sup>-1</sup> )	13.4	24.8	37.1	40.0***	26.1	21.9	19.1	21.9±1.6	1.4
Cu	(µg l <sup>-1</sup> )	4.5	2.9	5.9	9.1***	3.5	3.7	5.2	4.2±0.2	3.6*
Fe	(µg l <sup>-1</sup> )	1.5	0.8	2.0	3.0*	1.3	0.9	1.6	1.3±0.1	1.1
Cd	(µg l <sup>-1</sup> )	3.2	1.4	1.5	8.7***	2.3	2.1	2.4	2.3±0.2	0.2
Pb	(µg l <sup>-1</sup> )	1.7	0.9	2.4	4.9**	1.0	1.5	1.9	1.5±0.1	2.2
Zn	(µg l <sup>-1</sup> )	7.1	3.8	2.8	16.5***	6.0	4.8	4.9	5.2±0.4	0.8
Cl	$(\mathbf{gm} \ \mathbf{l}^{\cdot 1})$	3.1	0.9	0.7	25.5***	2.2	1.7	1.9	1.9±0.2	0.3

 Table (4): Pearson correlation coefficient (r-values) between the water variables and that of *P. australis* in Lake Burullus.\*;P
 0.05,

 \*\*;P
 0.01, \*\*\*;P
 0.001.

			Water				Leaf					Stem			Rhizon	ne
		Fe	Pb	Zn	Cu	Fe	Pb	Zn	Cu	Fe	Cd	Pb	Zn	Cu	Fe	Pb
Water	Cu	0.93**	0.97***	-0.26	0.59	-0.11	0.78	0.32	0.67	0.76	-0.33	0.93**	0.19	-0.09	0.87**	0.98***
	Fe		0.84*	-0.13	0.83*	-0.01	0.67	0.56	0.84*	0.89**	-0.37	0.82*	0.42	0.10	0.93**	0.89**
	$\mathbf{C}\mathbf{d}$		0.02	0.94**	0.23	0.93**	-0.39	0.55	0.57	0.06	0.79	-0.17	0.76	0.61	0.16	0.09
	Pb			-0.31	0.41	-0.11	0.82*	0.13	0.56	0.61	-0.24	0.94**	0.06	-0.07	0.82*	0.98***
	Zn				0.14	0.93**	-0.60	0.50	0.41	-0.13	0.82*	-0.44	0.72	0.68	-0.05	-0.22
Leaf	Cu					0.16	0.22	0.87*	0.86*	0.92*	-0.33	0.37	0.72	0.28	0.70	0.50
	Fe						-0.49	0.48	0.51	-0.04	0.87*	-0.32	0.77	0.83*	0.09	-0.04
	Cd						-0.49	0.78	0.63	0.30	0.39	-0.37	0.91	0.80*	0.18	-0.15
	Pb							-0.23	0.27	0.37	-0.49	0.96**	-0.35	-0.25	0.75	0.83*
	Zn								0.79	0.76	0.00	-0.01	0.92**	0.37	0.38	0.21
Stem	Cu									0.75	0.10	0.47	0.80*	0.55	0.83*	0.66
	Pb												-0.13	-0.19	0.84*	0.95**
Rhizor	Fe															0.90**





the nutritive value in the different organs of P. australis in Lake Burullus. DE: digestible energy, ME: metabolized energy, NE: net energy and GE: gross energy.

and *Nuphar lutea* (L.) Sm. and *Potamogeton nodosusPoir*. (Mazej and Germ, 2009) seem to be good retaining filters for several heavy metals.

In addition the positive significant linear correlation between the concentrations of Pb (r = 0.98, P < 0.001) and Fe (r = 0.93, P < 0.01) in the studied plant and in the water indicates bioaccumulation. Eid *et al.* (2012) in the Lake Burullus and Demirezen and Aksoy (2004) in Sultan Marsh reported similar results for Typha.

Therefore, this study suggests that *P. australis* may be used and served as a good bioaccumulator and bioindicator of heavy metals in the underground parts at

low metal concentrations in natural and polluted brackish water conditions. Moreover, these positive linear correlations were found between the heavy metal concentrations in all plant organs and those in water, thus indicating the potential use of such organs for monitoring of the polluted water and advantages of using plant species as bio monitors. Our results showed that the underground organs were the primary areas of heavy metal accumulation. In particular, heavy metal concentrations in plant organs decreased in the order of rhizome > stem > leaf and could be are comparable with the study of (Weis *et al.* 2004).

The roots and rhizomes of *P. australis* can accumulate great amount of heavy metals because it has parenchyma with large intercellular air spaces in the cortex (Sawidis *et al.*, 1995). Also the roots may be regarded as a heavy metal filter, slowly accumulating these heavy metals and thus partly preventing their transport to aboveground biomass (Tayler, 1971 and Liu *et al.*, 2007).

All three organs showed significant differences in concentration of Cu, Fe, Pb, Cd and Zn, and suggesting low mobility from roots to rhizomes and to aboveground organs. Although the organs followed different decreasing trends of element concentration, the trend Fe > Zn > Mn > Pb > Cu > Cd > Co was found in each plant organ.

This trend was comparable with the results of Bonanno

and Giudice (2010).

The high concentration of Mg, Mn, Co, Fe and Cd in the rhizome of the common reed in the present study may be due to the presence of an iron plaque around the roots (St-Cyr and Crowder, 1988 and Eid *et al.*, 2010a,b). Iron plaque is commonly observed on wetland plant roots and may play a key role to detoxify pollutants and purify water in wetlands constructed (Wang and Peverly, 1996).

This plaque may protect the roots from heavy metals toxicity, probably by co-precipitation or adsorption of toxic metals (Ali *et al.*, 2002). More studies in the natural and created wetlands indicated that the majority of wetland plants retain higher amounts of heavy metals and nutrients in their roots than shoots (Windom *et al.*, 1976; Breteler and Teal, 1981; Keller *et al.*, 1998 and Bonanno and Giudice, 2010).

Wetland plants play an important role in the nutrient cycling due to uptake, storage and release processes. *P. australis* is used as a 'nutrient remover' in wetlands, especially those designed and constructed for waste water treatment and extract large amounts of mineral nutrients to reduce its content of domestic, industrial and agricultural wastewater (Reddy and Smith, 1987; Hammer, 1989; Cooper and Findlater, 1990 and Verhoeven and Van der Toorn, 1990).

 Table (5): Spatial variation in the mean concentration ± standard deviation of the organic contents (%) in *Phragmites australis* in Lake Burullus. TA: total ash, Car: total carbohydrates, TL: total lipids, CF: crude fibres, CP: crude protein,

 DCP: digestible crude protein and TDN: total digestible nutrients. \*:P
 0.05, \*\*:P
 0.01, \*\*\*:P
 0.001.

	Section (Sc)	Site (St)	TA	Car	TL	CF	СР	DCP	TDN
	E-W	West	14.16±1.35	$40.81 \pm 4.06$	$2.78\pm0.42$	$28.58 \pm 1.56$	13.68±3.47	9.19±3.22	42.11±2.21
		Middle	$14.19 \pm 1.50$	$38.32 \pm 2.12$	$2.95\pm0.31$	$28.56 \pm 1.19$	$15.88 \pm 2.33$	$11.33 \pm 2.21$	$40.69 \pm 1.62$
		East	$14.11 \pm 1.54$	$40.79 \pm 3.32$	$2.88\pm0.35$	28.35±1.37	$13.82 \pm 3.31$	$9.32 \pm 3.07$	42.12±2.15
	N-S	South	$14.04 \pm 1.43$	$40.46 \pm 3.85$	$2.77 \pm 0.41$	$28.14{\pm}1.36$	$14.58 \pm 3.43$	$10.03 \pm 3.19$	$41.49 \pm 2.20$
Leaf		Middle	$14.13 \pm 1.51$	$40.47 \pm 2.97$	$2.96\pm0.29$	28.54±1.33	$13.85 \pm 3.08$	$9.34 \pm 2.86$	$42.16 \pm 2.01$
		North	$14.28 \pm 1.50$	$38.88 \pm 2.84$	$2.93\pm0.33$	$28.79 \pm 1.28$	$15.01 \pm 2.94$	$10.52 \pm 2.79$	$41.28 \pm 1.99$
	Total		14.15±1.43	39.87±3.28	$2.88 \pm 0.35$	28.48±1.31	14.56±3.10	$10.04 \pm 2.91$	41.58±2.04
	F-value	Sc	0.5	16.6***	1.6	0.2	29.7***	36.2***	25.8***
		St	4.0*	9.0**	2.1	1.6	11.7***	14.8***	12.0***
		Sc*St	7.6	4.2*	1.8	0.6	8.7***	10.2***	8.9***
	E-W	West	4.69±1.31	$53.73 \pm 5.50$	$0.88 \pm 0.27$	$35.45 \pm 5.05$	$5.29 \pm 1.04$	$2.56 \pm 4.16$	46.84±0.79
		Middle	$4.93 \pm 1.34$	$52.93 \pm 4.77$	$0.95 \pm 0.24$	$36.28 \pm 3.89$	4.91±0.46	$1.04\pm0.43$	$47.24\pm0.47$
		East	4.93±1.65	$54.19 \pm 5.62$	0.75±0.13	$35.28 \pm 4.55$	$4.86\pm0.57$	$0.98\pm0.54$	47.07±0.34
	N-S	South	$4.71 \pm 1.40$	53.11±4.85	$0.86\pm0.26$	$35.73 \pm 4.52$	$5.60 \pm 1.26$	$2.46 \pm 3.41$	$46.62 \pm 0.87$
<u>Stem</u>		Middle	$5.08 \pm 1.38$	$51.94 \pm 5.52$	$0.98\pm0.21$	$36.28\pm5.20$	$5.85 \pm 2.11$	$1.91 \pm 1.96$	$46.52 \pm 1.48$
		North	$4.94{\pm}1.39$	$53.56 \pm 5.10$	$0.93 \pm 0.21$	$35.85 \pm 4.26$	4.75±0.49	$0.89 \pm 0.46$	47.34±0.46
	Total		4.89±1.35	52.99±4.99	0.91±0.23	35.91±4.46	5.34±1.37	1.73±2.36	46.86±1.00
	F-value	Sc	6.9**	1.9	0.6	1.7	1.8	1.2	2.3
		St	7.7**	2.4	1.2	0.3	2.7	1.9	3.2
		Sc*St	2.5	5.4**	5.8**	8.1**	0.3	0.7	0.1
	E-W	West	7.63±0.71	$53.59 \pm 1.60$	$3.90\pm0.80$	$30.88 \pm 0.97$	$4.02\pm0.21$	0.21±0.20	49.63±0.58
		Middle	6.85±0.90	53.22±0.92	$3.94 \pm 0.78$	32.03±0.83	3.91±0.22	$0.16\pm0.14$	49.70±0.54
		East	$7.05\pm0.81$	53.14±1.17	3.95±0.84	31.84±0.86	4.96±3.13	$0.27 \pm 0.20$	49.73±0.56
Rhizome	N-S	South	7.16±0.92	53.31±1.20	$3.82 \pm 0.79$	31.67±0.97	4.05±0.18	$0.24\pm0.16$	49.55±0.56
		Middle	6.79±0.69	$53.34 \pm 0.87$	$3.99 \pm 0.74$	32.01±0.92	$3.88 \pm 0.32$	$0.24\pm0.17$	$49.78 \pm 0.46$
		North	7.13±0.93	53.12±1.23	4.01±0.85	31.71±0.94	$4.90 \pm 3.14$	0.17±0.19	49.77±0.57
	Total		7.05±0.85	53.24±1.09	3.94±0.77	31.78±0.92	4.33±1.95	0.21±0.17	49.70±0.53
	F-value	Sc	14.7***	0.7	0.1	2.8	0.6	1.4	0.4
		St	1.9	0.4	1.7	0.0	0.5	0.8	3.3
		Sc*St	4.5*	4.3*	5.6*	0.6	0.6	0.8	2.3

**Table (6):** Spatial variation in the mean concentration  $\pm$  standard deviation of the nutritive value (Mcal kg<sup>-1</sup>) of the individuals of*P. australis* in Lake Burullus in the north Nile Delta region. DE: digestible energy, ME: metabolized energy NE: net energyand GE: gross energy. \*;P0.05, \*\*:P0.01, \*\*\*;P0.001.

	Section (Sc)	Site (St)	DE	ME	NE	GE
	E-W	West	1.70±0.37	1.40±0.31	0.70±0.15	40.59±1.39
		Middle	$1.86\pm0.14$	$1.53\pm0.12$	$0.76\pm0.06$	$41.07 \pm 1.07$
		East	$1.77 \pm 0.17$	$1.45\pm0.14$	$0.73 \pm 0.07$	40.66±1.43
	N-S	South	$1.72\pm0.32$	$1.41\pm0.26$	0.70±0.13	40.76±1.37
<u>Leaf</u>		Middle	$1.80\pm0.14$	$1.48\pm0.12$	$0.74 \pm 0.06$	40.71±1.35
		North	$1.85 \pm 0.15$	$1.52\pm0.12$	$0.76 \pm 0.06$	$40.89 \pm 1.22$
			1.79±0.23	1.47±0.19	0.73±0.09	40.80±1.27
	Total E value	Sc	2.6	2.6	2.6	21.7***
	r-value	St	2.3	2.3	2.3	5.9**
		Sc*St	0.7	0.7	0.7	9.9***
	E-W	West	$1.45 \pm 0.07$	$1.19\pm0.05$	$0.60\pm0.03$	$42.50\pm0.42$
		Middle	$1.49 \pm 0.07$	$1.22\pm0.06$	$0.61 \pm 0.03$	42.95±1.36
		East	$1.46\pm0.03$	$1.19\pm0.03$	$0.60\pm0.01$	42.23±0.45
<u>Stem</u>	N-S	South	$1.48 \pm 0.07$	$1.22\pm0.06$	$0.61 \pm 0.03$	42.54±0.51
		Middle	$1.52 \pm 0.11$	$1.25\pm0.09$	$0.63 \pm 0.05$	42.60±0.54
		North	1.48±0.03	1.22±0.03	0.61±0.01	42.83±1.39
			1.49±0.07	$1.22\pm0.1$	$0.61 \pm 0.0$	42.66±0.93
	Total E volue	Sc	4.6*	4.8*	4.8*	17.9***
	r-value	St	0.9	1.0	1.0	3.5
		Sc*St	6.4**	6.6**	6.7**	14.7***
	E-W	West	$1.55 \pm 0.08$	$1.27 \pm 0.07$	0.64±0.03	42.36±0.65
		Middle	$1.62\pm0.05$	1.33±0.04	$0.66 \pm 0.02$	42.80±0.73
		East	$1.60\pm0.06$	$1.31\pm0.04$	$0.66 \pm 0.02$	42.74±0.70
<b>Rhizome</b>	N-S	South	$1.58 \pm 0.07$	$1.30\pm0.06$	$0.65 \pm 0.03$	42.59±0.77
		Middle	$1.62\pm0.06$	$1.33\pm0.05$	$0.67 \pm 0.02$	42.83±0.68
		North	$1.60 \pm 0.05$	1.31±0.04	$0.65 \pm 0.02$	42.73±0.70
			$1.60 \pm 0.06$	$1.31 \pm 0.05$	0.66±0.03	42.71±0.70
	Total	Sc	12.2***	13.6***	4.9*	17.9***
	F-value	St	3.3	3.7	0.3	3.5
		Sc*St	13.2***	14.2***	6.0**	14.7***

Table (7): Comparison between the nutrient and heavy metal
contents in the tissues of P. australis as estimated in the
present study with that of Eid et al., (2012).

**Table (8)**: Comparison between the organic contents and nutritive values of *P. australis* shoot in the present study and that of (El-Kady, 2000).

Element	Present study		Ei	Eid et al. (2012)					
	leaf	stem	rhizome	leaf	stem	rhizome			
Heavy me	Heavy metals (µg g <sup>-1</sup> dry wt)								
Pb	35.0	67.1	95.5	2202.5	2149.0	2163.5			
Cu	33.2	26.7	22.7	36.9	40.2	56.9			
Cd	18.8	16.4	23.4	15.3	14.8	14.7			
Zn	291.3	294.8	306.5	56.9	49.1	74.8			
Fe	823.5	824.3	2134.4	1046.0	1078.0	2045.5			
Ash(%)	14.2	4.9	7.1	13.1	7.4	5.7			
Nutrients	(mg g <sup>-1</sup> c	lry wt)							
Na	7.8	8.3	7.1	3.2	3.8	8.0			
К	13.3	13.7	11.6	10.8	7.8	7.3			
Ca	8.1	8.5	7.2	5.1	2.8	7.0			
Mg	5.9	6.6	5.4	3.6	1.4	2.9			
Р	0.2	0.2	0.1	1.0	0.4	0.5			
Ν	2.3	0.9	0.6	14.8	5.8	7.3			

Contents	Present study	El-Kady (2000)
Organic contents (%)	)	
Total ash	9.55	11.0
Total carbohydrates	46.45	50.6
Total lipids	1.9	1.9
Crude fibers	32.2	29.9
Crude protein	9.95	6.7
Digestible crude	5.85	2.9
protein Total digestible nutrients	46.0	31.4
Nutritive value (Mca	l/kg)	
Digestible energy	1.65	2.5
Metabolized energy	1.35	2.0
Net energy	0.65	1.0
Gross energy	41.7	40.3

**Table (9):** Comparison between the total digestible nutrients (TDN) values of *P. australis* in the present study and other related studies of fodder species.

Species	Study	TDN(%)
Phragmites australis	Present study	46.00
Panicum turgidum	Heneidy(1966)	64.03
Fodder crops:	Soliman & El-Shazly (1978)	
Clover (Trifolium alexandrinu)	· · ·	56.00
Barley (Hordeum vulgar)		64.00
Corn (Zea mayz)		68.00

 Table (10): Comparison of *P. australis* forage quality according to Boudet & Riviere (1968).

Nutritivecomparing	Net energy	Digestible
	(MJ kg <sup>-1</sup> )	protein (%)
Present Study:	4.18	5.85
Forage quality of Boudet		
& Riviere (1968):		
Poor	< 3.10	< 2.5
Fair	3.10 - 3.45	2.5 - 3.4
Good	3.45 - 4.15	3.4 - 5.3
Excellent	> 4.15	> 5.3

Several authors showed that production and nutrient storage in shoots and roots of *Phragmites* increase when nutrient availability increases (Mason and Bryant, 1975 and Ulrich and Burton, 1985). The results of the present study indicates the concentrations of nutrients are decreased in the order of stems > leaves > rhizomes. This trend is in accordance with the study of Dinka, (1986) and Eid *et al.*, (2012), whereas the underground organs are the most significant accumulators of all heavy metals (Németh and Lakner, 2002).

It is of interest to compare the heavy metals and nutrients in the present study with that of Eid, 2012 on the same plant in the same area (Table 7). The difference between the two studies could be interpreted in the view that the present study being mainly concerned with the spatial variations (E-W and N-S sections); while the past study dealt with temporal variations. Generally, it seems that heavy metal contents of *Phragmites* in Lake Burullus are higher than the corresponding range found in other studies such as in Egypt (El-Kady, 2000), Hungary (Engloner, 2004) and Turkey (Duman *et al.*, 2007).

This might be related to the pollution in Lake Burullus as a result of the input of all domestic, industrial and agri cultural wastes from the reclaimed lands that surround the lake. Furthermore, most of the estimated heavy metal and nutrients increased from north to south due to the accumulation of sewage effluents from the drains at the south of the lake, which agrees with the study of Shaltout *et al.*, (2004) and Eid *et al.*, (2012).

In the present study, the ash content accumulated in the leaves and rhizomes of population of common reed exhibited high annual mean. This may be due to the higher inorganic elements, which correlated with the ash contents (Ksenofontova, 1988). Thus, the ash value can be taken to give an approximate estimate of total inorganic residue of the plant. The ash content ranged from 4.9 to 14.2 % dry weight, which is comparable to the values reported by Ho. (1981); Best and Dassen (1987); Batanouny *et al.*, (1991) and El-Kady, 2000 (Table 8). The forage value of the consumed plant is the result of its nutritive value, i.e. chemical composition and digestibility. It is highly affected by the stage of maturity, edaphic influence; climate and range condition (Le Houérou, 1980). In addition, the nutritive value of any forage is dependent upon its content of energyproducing nutrients as well as its contents of essential elements to the body.

In the present study, the forage value of P. australis was evaluated according to its chemical composition where it indicated that Phragmites is a high-quality forage for domestic animals (such as sheep, goats and cattle), especially young leaves which have the highest percentages of total ash, crude protein and digestible crude protein; and young stems which attain the highest concentrations of total carbohydrates and crude fibers as compared with the same plant in the water courses of Nile Delta (Table 8: El-Kady, 2000). The range of these organic contents was in accordance with that reported by Duke (1998) who indicated that this plant contains 11.4% protein, 2.3% fat, 42.1% carbohydrates, 31.1% crude fiber and 10.8% ash, and hence the common reed is high-quality forage for cattle and horses and may be cut for hay.

The quality of forage can be expressed in several parameters, such as total digestible nutrients (TDN), digestible crude protein and caloric value (Duivenbooden, 1985). The total digestible nutrients are only an appropriate measure of the food energy available to animals after the digestion losses have been deduced (Lofgreen, 1951).

The value of TDN of Phragmites, as indicated in the present study, ranged from 41.6 % in leaf to 49.7 % in the rhizome which was lower than that of some grazed wild plants such as *Panicum turgidum* (Heneidy, 1996) and cultivated common fodder crops such as clover, barley and corn (Soliman and El-Shazly, 1978). However, it approximates the diet requirements for sheep (61.7%; NRC, 1975) and breeding cattle (50.0%; NRC, 1984a) (Table 9).

The Ministry of Agriculture, Fisheries and Food in England (1975) reported that the minimum proteins in the animal diet range between 6 and 12% depending on the animal type. The present study indicates that the protein content of *P. australis* approaches the requirements for the animal diet. Low protein levels efficiency is associated with a relatively low voluntary feed consumption with protein deficient diet.

The metabolism of the rumen microbiota may be depressed by a deficiency in rumen nitrogen. This limitation will retard the rate of removal of organic matter from the rumen which, in turn, may reduce the intake. Low protein levels will affect the wool growth, which is determined by protein absorbed in the intestine, which in turn lends on ingested nitrogen sources (see El-Kady, 1987). The Ministry of Agriculture, Fisheries and Food in England (1975) also reported that the digestible energy should be about 5.4% and the protein requirement is about 4.44 % of the weight. In the present study, protein contents ranged from 4.3 % in rhizomes to 14.6 % in leaves on the average, which is higher than the proper level.

All range nutritionists face the problem of determining the nutritive content of the diet of range animals. Grazing animals often select their forage from a complex mixture of plant species (Edlefsen *et al.*, 1960; Hosten, 2007; Shaltout *et al.*, 2008; Heneidy, 2012 and Tilley and John. 2012). Öelberg (1965) reported that the nutritive value of any forage is dependent upon its content of energy-producing nutrients as well as its content of nutrients essential to the body, normally protein, minerals and vitamins.

The nutritive value of range forage is influenced in a major way by maturity stage, edaphic influences, plant species, climate, animal class and range condition. It may be suggested that animals should be supplied with supplementary feed rich in protein, particularly during the growth and reproductive stage, in order to maximize their productivity.

In the present study, the forage quality according to Boudet and Riviere (1968) showed that the green parts of *P. australis* are ranked under excellent fodder quality (Table 10), where the net energy are (4.18 MJ kg<sup>-1</sup>) and digestible protein (5.85%). In general, the organic components and nutritive values of green parts are within the ranges in the feeds commonly used in ratios of sheep, goat and cattle (NRC, 1975; 1978; 1981 and 1984a, b).

### CONCLUSION

In general, as accumulation of the analyzed heavy metals was higher in underground organs, there seems to be no danger for animals to feed on the aboveground parts, which are rich in their nutritive value. Hence, the current results are sufficient to indicate that *P. australis* can be use as a potential biological barrier against the spread of heavy metal pollution and can be using as a fodder plant in lakes. Similar conclusion has been made for five lakes in Poland (Drzewiecka *et al.*, 2010).

#### REFERENCES

- ALI, N. A., M. P. BERNAL, AND M. ATER. 2002. Tolerance and bioaccumulation of copper in *Phragmites australis* and *Zea mays*. Plant and Soil **239**: 103 - 111.
- ALLEN, S., H. M. GRIMSHAW, J. A. PARKINSON, AND C. QUARMBY. 1989. Chemical Analysis of Ecological Materials. 4<sup>th</sup> ed., Blackwell Scientific Publications, London.
- ÁLVAREZ-ROGEL, J., F. J. JIMÉNEZ-CÁRCELES, AND C. EGEA. 2006. Phosphorus and Nitrogen Content in the Water of a Coastal Wetland in the Mar Menor Lagoon (SE Spain): Relationships with Effluents from Urban

and Agricultural Areas. Water, Air, and Soil Pollution **173**: 21 - 38.

- ANONYMOUS. 1980. Climatic Normals for the Arab Republic of Egypt up to 1975. Ministry of Civil Aviation, Meteorological Authority, General Organization for Governmental Printing Offices, Cairo.
- ASAEDA, T., L. NAM, P. HIETZ, N. TANAKA, AND S. KARUNARATNE. 2002. Seasonal Fluctuations in Live and Dead Biomass of *Phragmites australis* as Described by a Growth and Decomposition Model: Implications of Duration of Aerobic Conditions for Litter Mineralization and Sedimentation. Aquatic Botany **73**: 223 239.
- BALDATONI, D., A. ALTONI, P. D. TOMAMASI, B. GIOVANNI, DE. VIRZO, AND A. SANTO. 2003. Assessment of Macro and Microelement Accumulation Capability of two Aquatic Plants. Environmental Pollution **130**: 149 156.
- BATANOUNY, K. H., A. H. HASSAN, AND N. A. SAWAF. 1991. Ecotypes of *Phragmites australis* in Egypt. Proceeding International Conference: Plant Growth, Drought and Salinity in the Arab Region 97 - 114.
- BEST, E. P. H., AND J. H. A. DASSEN. 1987. A seasonal study of growth characteristics and the levels of carbohydrates and proteins in *Elodea nuttallii*, *Polygonum amphibium and Phragmites australis*. Aquatic Botany **28**: 353 372.
- BONANNO, G., AND R. L. O. GIUDICE. 2010. Heavy metal bioaccumulation by the organs of *Phragmites australis* common reed) and their potential use as contamination indicators. Ecological Indicators **103**: 639 645.
- BORIN, M., G. BONANITI, G. SANTAMARIA, AND L. GIARDINI. 2001. A Constructed Surface Flow Wetland for Treating Agricultural Waste Waters. Water Science and Technology **44**: 523 530.
- BOUDET, G., AND R. RIVIERE. 1968. Emploi pratique des analyses fourragères pour l'appréciation des pâturages tropicaux. Revue d'Elevage et de Médecine Vétérinaire des Pays Tropicaux **2** (21): 227 - 266.
- BREKKEN, A., AND E. STEINNES. 2004. Seasonal concentrations of cadmium and zinc in native pasture plants: consequences for grazing animals. Science of the Total Environment **326**: 181-195.
- BRETELER, R. J., AND J. M. TEAL. 1981. Trace element enrichments in decomposing litter of *Spartina alterniflora*. Aquatic Botany **11**: 111 - 120.
- BRIX, H., AND H. H. SCHIERUP. 1989. The Use of Macrophytes in Water-Pollution Control. Ambio 18: 100-107.
- CIRUJANO, S., M. MORENO, A. RUBIO, AND J. ECHEVERRÍAS. 2005. *Capacidad depuradora* del carrizo en el Parque Natural El Hondo (Alicante). Biodiversidad y Gestión de los carrizales. In: Actas de las I Jornadas Científicas Parque Natural de El Hondo, Biodiversidad y Gestión de los carrizales, Crevillente.
- COOPER, P. F., AND B. C. FINDLATER. 1990. Proceedings of the International Conference on the Use of

- Constructed wetlands in Water Pollution Control, **09**: 24 28.Cambridge, UK
- DEMARQUILLY, C., AND P. WEISS. 1970. Tableau de la valeur alimentaire des fourrages. Et. 42: Versailles INRA-SEI.
- DEMIREZEN, D., AND A. AKSOY. 2004. Accumulation of heavy metals in *Typha angustifolia L*. and *Potamogeton pectinatus L*. living in Sultan Marsh (Kayseri, Turkey). Chemosphere **56**: 685 - 696.
- DINKA, M. 1986. The effect of mineral nutrient enrichment of Lake Balaton on the common reed (*Phragmites australis*). Folia Geobotanica et Phytotaxonomica **21**: 65 - 84.
- DRZEWIECKA, K., K. BOROWIAK, M. M. LECZEK, AND I. ZAWADAI. 2010. Cadmium and lead accumulation in two littoral plants of five lakes in Poznan, Poland. Acta Biologica Cracovienisia Series Botanica **52**(2): 59 68.
- DUMAN, F., M. CICEK, AND G. SEZEN. 2007. Seasonal changes of metal accumulation and distribution in common club rush (*Schoenoplectus lacustris*) and common reed (*Phragmites australis*). Ecotoxicology **16** (6): 457 463.
- EDLEFSEN, J. L., C. W. COOK, AND I. T. BLACK. 1960. Nutrient content of the diet as determined by hand plucked and oesophageal fistula samples. Journal of Animal Science 9 (2): 2 - 20.
- EID, E. M., K. H. SHALTOUT, Y. M. AL-SODANY, AND K. JENSEN. 2010a. Effects of abiotic conditions on *Phragmites australis* along geographic gradients in Lake Burullus, Egypt. Aquatic Botany **92**: 86-92.
- EID, E. M., K. H. SHALTOUT, Y. M. AL-SODANY, K. SOETAERTAND, AND K. JENSEN. 2010b. Modeling growth, carbon allocation and nutrient budget of *Phragmites australis* in Lake Burullus, Egypt.Wetlands **30**: 240 - 251
- EID, E. M., M. A. EL-SHEIKH, AND A. A. ALATAR. 2012. Uptake of Ag, Co and Ni by the Organs of Typha domingensis (Pers.) Poir. ex Steud. in Lake Burullus and Their Potential Use as Contamination Indicators. Open Journal of Modern Hydrology **02**: 21 - 27.
- EID, E. M. 2012. *Phragmites australis* (Cav.) Trin. ex Steud.: Its Population Biology and Nutrient Cycle in Lake Burullus, a Ramsar site in Egypt. LAP LAMBERT Academic Publishing, Saarbrücken, Germany, pp. 292.
- EL-KADY, H. F. 1987. A study of Range Ecosystems of the Western Mediterranean Coast of Egypt. Ph.D. Thesis, Technical University, Berlin.
- EL-KADY, H. F. 2000. Seasonal variation in phytomass and nutrient status of *Phragmites australis* along the water courses in the Middle Delta Region. Taeckholmia **20** (2): 123 - 138.
- ENGLONER, A. 2004. Annual growth dynamics and morphological differences of reed *Phragmites australis* (Cav.) Trin. ex Steudel in relation to water supply. Flora **199**: 256 - 262.
- FRANKENBER, J. 1997. Guidelines for growing Phragmites for erosion control. Cooperative Research

Centre for Freshwater Ecology. Murray-Darling Freshwater Research Centre. Albury, NSW, Australia.

- GABRA, M. A., N. M. EL-KHOLY, S. Y. SHERIF, AND I. M. SOLIMAN. 1987. Chemical composition and feeding value of fooder beet (*Beta vulgaris L*.). 1st Proceeding Conference Development Research, Faculty Agriculture, Ain Shams University Egypt **1**: 92 - 100.
- GARRETT, W. N. 1980. Energy Utilization of Growing Cattle as Determined In Seventy-Two Comparative Slaughter Experiments," in Energy Metabolism. L. E. Mount, [Ed.], vol. **26**, EAAP, London, UK.
- GÓMEZ, C. R., M. L .SUÁREZ, AND M. R. VIDAL-ABARCA. 2000. The Performance of Multi-stage System of Constructed Wetlands for Urban Wastewater treatment in a Semiarid Region of SE Spain. Ecological Engineering **16**: 501 - 517.
- GREENBERG, A. E., L. S. CLESCERI, AND A. D. EATON. 1992. Standard Methods for the Examination of Water and Wastewater. American Public Health Association, Washington, DC.
- HAMMER, D. A. 1989. Constructed Wetlands for Wastewater Treatment: Municipal, Industrial and Agricultural. Lewis Publishers Inc., Chelsea, MI, USA.
- HENEIDY, S. Z. 1996. Palatability and nutritive values of some common plant species from the Aqaba Gulf area of Sinai, Egypt. Journal of Arid Environments **34**: 115 123.
- HENEIDY, S. Z. 2012. Rangelands in Arid Ecosystem, Diversity of Ecosystems. Mahamane Ali, [Ed.], ISBN: 978-953-51-0572-5, InTech, Available from: <u>http://www.intechopen.com/books/diversityof</u> ecosystems/rangelands-in-arid-ecosystem
- Ho, Y. B. 1981. MINERAL composition of *Phragmites australis* in Scottish lochs as related to eutrophication I: seasonal changes in organs. Hydrobiologia **85**: 227 237.
- HOLM, L. G., D. L. PLUCKNETT, J. V. PANCHO, AND J. P. HERBERGER. 1977. *Phragmites australis* (Cav.) Trin. (= *P. communis* Trin.) and *Phragmites karka* (Retz.) Trin. In: The World's Worst Weeds "Distribution and Biology". The University Press of Hawaii, Honolulu.
- HOSTEN, P. E., H., WHITRIDGE, D. SCHUSTER, AND J.ALEXANDER. 2007. Livestock on the Cascade-Siskiyou National Monument: A Summary of Stocking Rates, Utilization, and Management. U.S. Department of the Interior, Bureau of Land Management, Medford District. http://soda.sou.edu/bioregion.html
- KELLER, B., K. LAJTHA, AND S. CRISTOFOR. 1998. Trace metal concentrations in the sediments and plants of the Danube delta, Romania. Wetlands **18**: 42 50.
- KIEDRZYNSKA, E., I. WAGNER, AND M.ZALEWSKI. 2008. Quantification of phosphorus retention efficiency by floodplain vegetation and a management strategy for a eutrophic reservoir restoration. Ecological Engineering **33**: 15 - 25.
- KSENOFONTOVA, T. 1988. Morphology, production and mineral contents in *Phragmites australis* in different

waterbodies of the Estonian SSR. Folia Geobotanica et Phytotaxonomica **23**: 17 - 43.

- LE HOUÉROU, H. N. 1980. Chemical Composition and Nutritive Value of Browse in Tropical West Africa. In H. N. Le Hou´erou, [ed.], Browse in Africa, ILCA, Addis Ababa, Ethiopia, pp. 261 - 289.
- LELONG, B., C. LAVOIE, Y. JODOIN, AND F. BELZILE. 2007. Expansion Pathways of the Exotic Common reed (*Phragmites australis*): A Historical and Genetic Analysis. Diversity and Distributions **13**: 430 - 437.
- LEWANDER, M., M. GREGER, I. KAUTSKY, AND E. SZAREK. 1996. Macrophytes as indicators of bioavailable Cd, Pb and Zn flow in the river Przemsza, Katowice Region. Applied Geochemistry **11**: 169 173.
- LIU, J., Y., DONG, H., XU, D. WANG, AND J. XU. 2007. Accumulation of Cd, Pb and Zn by 19 wetland plant species in constructed wetland. Journal of Hazardous Materials **147**: 947 - 953.
- LOFGREEN, G. P. 1951. The use of digestible energy in evaluation of feeds. Journal of Animal Science **10**: 344 351.
- MARTIN, M. AND P. COUPHTREY. 1982. Biological Monitoring of Heavy Metal Pollution. Applied Sciences Publications, London/New York.
- MASON, C. F. AND R. J. BRYANT. 1975. Production, nutrient content and decomposition of *Phragmites communis* Trin. and *Typha angustifolia* L. Journal of Ecology **63** (1): 71 - 95.
- MAZEJ, Z., AND M. GERM. 2009. Trace element accumulation and distribution in four aquatic macrophytes. Chemosphere **74**: 642 647.
- MEULEMAN, A., J. BEEKMAN, AND J. VERHOEVEN. 2002. Nutrient retention and nutrient-use efficiency in *Phragmites australis* Stands after wastewater application. Wetlands **22**: 712 - 721.
- MINCHINTON, T. E., AND M. D. BERTNESS. 2003. Disturbance mediated competition and the spread of *Phragmites australis* in a coastal marsh. Ecological Applications **13**: 1400 - 1416.
- Ministry of Agriculture Fisheries and Food in England. 1975. Energy allowances and feeding system for ruminants. Technical Bulletin 33, Her Majesty's Station Office, London.
- NAGA, M. A. AND K. EL-SHAZLY. 1971. The prediction of the nutritive value of animal feeds from chemical analysis. Journal of Agricultural Science **77**: 1 - 25.
- NÉMETH, N. AND G. LAKNER. 2002. Some aspects of the purification efficiency of *Phragmites australis* in a root zone system. Acta Biologica Szegediensis
- NEWMAN, S., AND K. PIETRO. 2001. Phosphorus storage and release in response to flooding: implications for everglades stormwater treatment areas. Ecological Engineering **18**: 23 - 28.
- NRC. 1975. Nutrient Requirements of Domestic Animals: Nutrient Requirement of Sheep 5<sup>th</sup> [ed.], National Research Council No. 5, National Academy of Science, Washington DC, USA.
  - NRC. 1978. Nutrient Requirements of Domestic Animals: nutrient requirement of dairy cattle 5<sup>th</sup>

[ed.], National Research Council No. 3, National Academy of Science, Washington, DC, USA.

- NRC (1981). Nutrient requirements of domestic animals: nutrient requirement of goats. National Research Council No. 15, National Academy of Science, Washington, DC, USA.
- NRC. 1984a. Nutrient Requirements of Domestic Animals: Nutrient Requirement of Beef Cattle. 6<sup>th</sup> [ed.], National Research Council No. 5, National Academy of Science, Washington, DC, USA.
- NRC. 1984b. Nutrient Requirements of Domestic Animals: Nutrient Requirement of Sheep. National Research Council no. 5, 6<sup>th</sup> [ed.], National Academy of Sciences, Washington, DC, USA.
- ÖELBERG, K. 1965. Factors affecting the nutritive value of range forages. Journal of range Management **9**: 220 225.
- PEVERLY, J. H., J. M. SURFACE, AND T. WANG. 1995. Growth and trace metal absorption by *Phragmites australis* in wetlands constructed for landfill leachate treatment. Ecological Engineering **5**: 21-35.
- POULIN, B., G. LEFEBVRE, AND A. MAUCHAMP. 2002. Habitat requirements of passerines and reedbed management in Southern France. Biological Conservation **107**: 315 - 325.
- RASKIN, I., U. KRAMER, R. D. SMITH, D. E. SALT, AND R. SCHULMAN. 1997. Phytoremediation and mechanisms of metal accumulation in plants. Plant Physiology 114: 1253 - 1255.
- REDDY, K. R. AND W. H. SMITH. 1987. Aquatic Plants for Wastewater Treatment and Resource Recovery. Magnolia Publishing Inc., Orlando, Florida, USA.
- RUIZ, M. AND J. VELASCO. 2010. Nutrient bioaccumulation in *Phragmites australis*: management tool for reduction of pollution in the Mar Menor. Water Air Soil Pollution **205**: 173 185.
- SALT, D. E., R .D. SMITH, AND I. RASKIN. 1998. Phytoremediation. Annual Review of Plant Physiology **49**: 643 - 668.
- SAWIDIS, T., M. K., CHETTRI, G. A. ZACHARIADIS, AND J. A. STRATIS. 1995. Heavy metals in aquatic plants and sediments from water systems in Macedonia, Greece. Ecotoxicity Environment Safety 32: 73 - 80.
- SHALTOUT, K. H., Y. M. AL-SODANY, AND M. A. EL-SHEIKH. 2004. *Phragmites australis* (Cav.) Trin. ex Steud. In Lake Burullus, Egypt: is it an expanding or retreating population. The Third International Conference on Biological Sciences **3**: 83 - 96.
- SHALTOUT, K. H. AND M. T. KHALIL. 2005. Lake Burullus: Burullus Protected Area. Publication of National Biodiversity Unit No. 13, Egyptian Environmental Affairs Agency (EEAA), Cairo.
- SHALTOUT, K. H., AND Y. M. AL-SODANY. 2008. vegetation analysis of Burullus wetland: A RAMSAR site in Egypt. Wetland and Management Ecology **16**: 421 439.
- SHALTOUT, K. H., A. A. EL-KEBLAWY, AND M. T. MOUSA. 2008. Evaluation of the range plants quality and palatability for camel grazing in the United Arab Emirates. Journal of Camelid Sciences 1: 1 13.

- SOLIMAN, S. M., AND K. EL-SHAZLY. 1978. Increasing the productivity per feddan from total digestible nutrients. Alexandria Journal of Agriculture Research **26**: 551 - 556.
- SOUTHICHAK, B., K. NAKANO, M. NOMURA, AND N. CHIBA. 2006. *Phragmites australis*: A novel bioabsorbent for the removal of heavy metals from aqueous solution. Water Research **40**: 2295 2302.
- SPSS. 2006. SPSS Base 15.0 User's Guide. SPSS Inc., Chicago, USA.
- ST-CYR, L., AND A. A. CROWDER, 1988. Iron oxide deposits on the roots of *Phragmites australis* related to the iron bound to carbonates in the soil. Journal of Plant Nutrition **11**: 1253 1261.
- STOLTZ, E. AND M. GREGER. 2002. Accumulation properties of As, Cd, Cu, Pb and Zn by four wetland plant species growing on submerged mine tailings. Environmental and Experimental Botany **47**: 271 280.
- TAYLER, G. 1971. Distribution and turnover of organic matter and minerals in a shore meadow ecosystem. Oikos **22**: 265 291.
- TERRY, N., S. V. SAMBUKUMAR, AND D. L. LEDUC. 2003. Biotechnological approaches for enhancing phytoremediation of heavy metals and metalloids. Acta Biotechnologica **23**: 281 288.
- TILLEY, D. J. AND L.S. JOHN. 2012. Plant Guide for common reed (*Phragmites australis*). USDA-Natural Resources Conservation Service, Aberdeen, ID Plant Materials Center. 83210 - 0296.
- ULRICH, K. E., AND T. M. BURTON. 1985. The effects of Nitrate, Phosphate, and Potassium fertilization on growth and nutrient uptake patterns of *Phragmites australis*. Aquatic Botany **21**: 53 - 62.
- UNESCO. 1977. Map of the World distribution of arid regions. MAB Technical Notes 7.
- VERHOEVEN, J. T., AND J. VAN, DER. TOORN. 1990. Management of marshes with special reference to wastewater treatment. In B.C. Patten [ed.], Wetlands

and Shallow Continental Water Bodies. Vol I. Academic Publishing, The Hague, the Netherlands. pp. 571 - 585.

- VYAMAZAL, J. 2002. The use of Sub-surface Constructed Wetlands for Wastewater Treatment in the Czech Republic: 10 years Experience. Ecological Engineering **18**: 663 - 646.
- VYMAZAL, J., J. ŠVEHLA, L. KROPFELOVA, AND V. CHRASTNY. 2007. Trace elements in *Phragmites australis* and *Phalaris arundinacea* growing in constructed and natural wetlands. Science of the Total Environment **380**: 154 162
- VYMAZAL, J., L. KROPFELOVA, J. ŠVEHLA, V. CHRASTNY. AND J. ŠTICHOVA. 2009. Trace elements in *Phragmites australis* growing in constructed wetlands for treatment of municipal wastewater. Ecological Engineering **35**: 303 309.
- WANG, T. G., AND J. H. PEVERLY. 1996. Oxidation states and fractionation of plaque iron on roots of common reeds. Soil Science Society of American Journal **60**: 323 - 329.
- WEIS, J. S., AND P. WEIS. 2003. Is the invasion of the common reed, *Phragmites australis*, into tidal marshes of the eastern US an ecological disaster? Marine Pollution Bulletin **46**: 816 820.
- WEIS, J. S., T. GLOVER, AND P. WEIS. 2004. Interactions of metals affect their distribution in tissues of *Phragmites australis*. Environmental Pollution **131**: 409-415.
- WILKINS, D. A. 1978. The measurement of tolerance to edaphic factors by means of root growth. New Phytologist **80**: 623 633.
- WINDOM, H., W. GARDNER, J. STEPHENS, AND F. TAYLOR. 1976. The role of methyl mercury production in the transfer of mercury in a salt marsh ecosystem. Estuarine and Coastal Marine Science **4**: 579 583.
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