BENTHIC HABITAT MAPPING USING REMOTE SENSING DATA AT HURGHADA REGION, RED SEA COAST, EGYPT

Mostafa Khaled^{1, 2, 3}*, Ahmad Obuid-Allah³, Frank Muller-Karger¹, Mahmoud Ahmed², Sameh El-Kafrawy² and Ali A. Thabet

¹Institute for Marine Remote Sensing, College of Marine Science, University of South Florida, St. Petersburg, USA

²Department of Marine Science, National Authority for Remote Sensing & Space Science, Cairo, Egypt

³Department of Zoology, Faculty of Science, Assiut University, Assiut, Egypt

Received: 17/8/2020 Accepted: 16/9/2020 Available Online: 1/12/2020

The present research was designed to focus on the utility of Landsat 8-OLI multispectral data for identifying and classifying benthic habitats mapping of the Red Sea after applying atmospheric and water-column corrections at Hurghada city. Atmospheric and water column corrections were applied to the imagery, making it an effective method for mapping benthic habitats. Water column correction was achieved by deriving absorption and backscattering coefficients for each band of the image of clear water pixels. An unsupervised classification (ISODATA) algorithm was applied to generating 22 class habitats. The supervised classification was performed using machine-learning algorithm a maximum likehood and reference points to produce 7 classes of benthic habitat as the following, coral reefs (dense and patch), sea weeds (macro-algae), sea grass (dense and patch), deep water (more than 20 m), shallow water (less than 20 m), sandy bottom (mainly consist of calcium carbonates and silicates) and rocky bottom. Sea weeds (Macroalgae) and deep water areas showed the highest producer's and user's accuracies, when compared to dense seagrass, mixed: seagrass/sand, and mixed: coral/sand areas. Based on 1050 reference points overall accuracy of the benthic habitat assessment is 66.7 percent, with an overall Kappa coefficient value of 0.611.

Keywords: Mapping; Benthic habitat; Remote sensing, Accuracy assessment.

1. INTRODUCTION

The Red Sea is one of the most important marine biodiversity repositories in the world. It has abundant and unique coastal and marine habitats including coral reefs, mangroves, sea weeds and sea grass beds which perform important biological, ecological, aesthetic, and economic functions. They provide key resources for coastal populations: food, shoreline protection and stabilization, and economic benefits from tourism development [1, 2]. The Red Sea is the habitat of over 1,000 invertebrate species, more than 1200 species of fishes, and 200 soft and hard corals. It is the world's northernmost tropical sea [3].

Hurghada was a fishing city and later, in the 1980s became the first tourist resort on the Egyptian Red Sea [4]. Its coast extends for about 46 km along the Red Sea, and it is mainly supported by tourism from water-based sports and activities [5]. Monitoring can provide valuable information on the benthic habitats condition derived from different data sources such as satellite images; field surveys; and local knowledge [6, 7]. Anthropogenic impacts on the seafloor alter benthic biodiversity [8], habitats [9], and modify ecosystem structures and functions [10].

Remote sensing is widely used to provide accurate and timely geospatial information describing changes in urban land use/land cover [11, 12]. The importance of remote sensing was emphasized as a 'unique view' of the spatial and temporal dynamics of the processes in urban growth and land use change [13-15]. Satellite remote-sensing techniques have been used in detecting, mapping benthic habitat and monitoring land cover change at various scales with useful results [16-18]. The combination of remote sensing with geographical information systems (GIS) and field data collection have been used to assess land cover change and mapping [19-22].

Satellite data has been suggested as a potential tool for monitoring coral-reef ecosystems with several researchers having **tested space-borne** sensor systems [6]. Spatial resolution of these systems ranges from 30 m for the Landsat Thematic Mapper (TM) to 4 m for the Ikonos multispectral data. Those evaluating the utility of the TM have mapped subtidal coastal habitats [23], delineated sand bottoms [24], digitally classified coral reef zones [25], evaluated the benthos [26], and performed time series analyses [15, 27]. Similarly, researchers have used SPOT (20 m) imagery to survey coral reef abundance [28] and apply a spatial statistical approach to multidate SPOT data to identify coral stress [29].

Coral reefs consist of a mosaic of fine-scale features between 1 to 5 m in size with complex optical signatures that blend as individual fields of view become larger [27]. The relatively coarse spatial resolutions of TM and SPOT may have a limited effectiveness in coral reef studies. Mumby *et al.* [16] noted that although TM and SPOT can detect benthic signals through clear water to a depth of approximately 25 m, the coarse spatial and spectral resolutions of these systems limit classification results to broad-based geomorphological information, rather than biotic assemblages. The same authors concluded that a pixel size of 3 to 4 m is probably optimal for surveying tropical marine environments. High resolution satellite images such as Ikonos, QuickBird and WorldView would be deemed appropriate for evaluating coral habitats. Recent studies concentrated on using

sensors with high spatial resolution such as Ikonos, QuickBird, and WorldView in shallow water mainly to classify and map coral reef communities. These previous sensors help to increase the ability to discriminate the benthic habitat into classes [30, 31]. The combination between satellite data and bathymetric data also assisted in increasing the accuracy of benthic habitat mapping [32]. Early Studies analyzing Ikonos data to assess coral reefs were applied [32, 33]. The previous authors concluded that Ikonos great than Landsat-7 in the benthic habitat classification. QuickBird may have significant potential for discriminating and mapping benthic habitats in tropical-coastal environments [34].

Finally, a benthic habitat can be defined as an area of anything associated with or occurring on the bottom of the deep water seabed [35]. **Brown** *et al.* [36] provided a comprehensive review of benthic habitat map types, data collection techniques, and habitat maps methodologies. Habitat-based approaches have been used for decades in landscape ecology [37, 38]. Due to the species environmental range preferences and requirements, many of these approaches focused on the structure and quantity of potential habitats, in addition to, the distribution of biological populations at the sampling time [39]. Habitat maps must be placed in context with the appropriate spatial, temporal, and thematic scales [40]. Scale is only briefly acknowledged in the extensive literature on benthic habitat mapping, often with little or no treatment of the role of spatial scale in the production of benthic maps and the interpretation of research results [42].

The main objective of the present research was to use Landsat 8-OLI multispectral data for identify and classify tropical-marine benthic habitats, after applying atmospheric and water-column corrections.

2. MATERIALS AND METHODS

2.1. Study area

Hurghada lies between 27° 25 to 27° 9 N and 33° 40 to 34° 8 E. It is the main touristic city of the Egyptian Red Sea which extends from El Gouna Resort to Magawish Resort (Figure 1). It covers approximately 46 km in length and 35 km in width .



Figure 1: Landsat image indicating the studied location and benthic habitat transects.

2.2. Satellite Data

One Landsat 8-OLI multispectral image was collected for the studied site on 10 July 2015, which downloaded from the United States Geological Survey (http://earthexplorer.usgs.gov/). The Landsat 8-OLI sensor system has 11 bands; each band had slightly different spectral ranges (Table 1). The first four bands can be used for shallow marine applications.

2.3. Image processing

Satellite image was georeferenced and geometrically corrected to match a WGS 84 datum (world geographic system), UTM (Universe Transverse Mercator) Projection with Zone 36 North using ground control points verification using Guno Trimble GPS. This process was completed by using ERDAS software package with Root Mean Square Error (RMSE) of geometric corrected 0. 0035. The image data were radiometrically, atmospherically and water column corrected is based on linear relationship of the spectral radiance and the reflectance of objects on the bottom of the shallow waters using the FLAASH atmospheric correction module of the ENVI® software package (**Figure 2**). Air-water interface correction was applied using coefficients extracted from **Andrefouet** *et al.* [33]. A manually digitized land

and deep-water mask, was applied to enable processing to focus on just the interand sub-tidal shallow waters.

Unsupervised classification (ISODATA) was performed on the image to yield **22** classes for water. The classes validation was conducted by selecting random points for each habitat in the area of the study around Hurghada. Ground truthing data collection was conducted over a period of **10 days in 2015**. About **150** reference points (**Figure 1**) were selected for each water classes in the study area using GPS points for the image data validation, data validation was gathered using georeferenced photo transects collected using snorkeling over the benthic habitat while taking photos with a digital housing camera at a set distance from the bottom, each photo was logged by Guno Trimble GPS and interpreted using Coral Point Count (**CPCe 4.0**) software. Locations of the logged photo were labeled based on common benthic habitat cover classes scheme (**Table 2**) similar to those defined by [43, 44].

A supervised classification and its accuracy assessment were then conducted by selecting specific training data for each benthic cover type. The supervised classification was performed using the model of machine-learning algorithm a maximum likehood and confusion matrix analyses (ENVI 5.2 software) for each image pixels, to calculate the overall accuracy of the classification result and kappa coefficient.



Figure 2: Landsat image before and after the atmospheric correction; (A) and (B) are the zoomed areas near Al Ahiaa district coral reef after and before the atmospheric correction.

Band Number	μm	Resolution	Band Number	μm	Resolution
1	0.433– 0.453	30 m	7	2.100– 2.300	30 m
2	0.450– 0.515	30 m	8	0.500– 0.680	15 m
3	0.525– 0.600	30 m	9	1.360– 1.390	30 m
4	0.630– 0.680	30 m	10	10.6-11.2	100 m
5	0.845– 0.885	30 m	11	11.5-12.5	100 m
6	1.560– 1.660	30 m			

 Table 1: Landsat 8-OLI sensor spectral bands.

No	Class	Туре	Description
1		Coral reef	Healthy, dead and infected coral reef included. Dominated by any life-form (digitate, branching, tabular, foliose, massive, submissive and encrusting) (>70%).
2		Seagras s	Seagrass bed mostly dense and patch dominated by <i>Halophila</i> sp. (>70%) and other species may be present in small quantity.
3		Sand bottom	Mainly calcium carbonate sand, white bright colour (>70%) and some small rubbles.
4		Sea weeds (Macro algae)	Area dominated by brown, green and mixed algae such as <i>Padina</i> sp., <i>Sargassum</i> sp., <i>Ulva</i> sp., <i>Caulerpa</i> sp., <i>Laurencia</i> sp., <i>Halimeda</i> sp., also turf brown algae (>70%).
5		Deep water	That's more than 20 m depth
6		Shallo w water	That's less than 20 m depth
7		Rock bottom	Area of mainly rubble (>70%), with small portion of macroalgae, seagrass, or dead coral present.

Table 2: Benthic habitats classification scheme used in the study.

3. RESULTS

Visually images before and after atmospheric correction revealed a sharp difference. The haziness depicted in the original Landsat image, attributed to Rayleigh and aerosol scattering, was eliminated resulting in a visually clear image as shown in the zoomed images (**Figure 2**). Radiance values of water over different bottom types were analyzed before and after the atmospheric correction.

A comparative evaluation of the classified image was performed against 1050 independents in situ points revealing an overall accuracy of 66.7 percent (**Table 3**). The overall Kappa statistic, a discrete multivariate accuracy assessment technique described by **Congalton and Mead [45]**, was **0.611 (Table 3**). This statistic estimates the percent of successful classifications compared to a random, chance classification assignment [46]. Sea weeds (Macroalgae) and deep water areas showed the highest producer's and user's accuracies, when compared to Dense seagrass, mixed: seagrass/sand, and Mixed: coral/sand areas. Sand (very bright), and deep water (very dark) are the two most spectrally distinct classes and yielded the lowest classification errors, whereas the mixed benthos areas had higher error because of the spectral similarities between various features.

The benthic habitat could be classified due to Maximum Likehood Algorithm and in situ data into seven classes as the following: coral reefs (dense and patch), sea weeds (macro-algae), sea grass (dense and patch), deep water (more than 20 m), shallow water (less than 20 m), sandy bottom (mainly consist of calcium carbonates and silicates) and rocky bottom according to scheme (**Table 2 Figure 3**).

Based on the accuracy assessment, difference in accuracy can be seen in each benthic habitat class according to producers and user's accuracy. Producer accuracy using Landsat 8-OLI in mapping 7 benthic habitat classes of Sea Weed (S.W.), Deep Water (D.W.) and Shallow Water (Sh.W.) are above 80% which are 92%, 86.67%, 85.33%; respectively. While the other classes are above 60% of Rocky Bottom (R.B.), Sea Grass (S. G.) and Coral reef (C.R.) which are 69.33%, 63.33%, 61.33%; respectively. Sandy Bottom (S.B.), show 8.67% (Table 3 and Figure 4). The results demonstrated that atmospheric and water column correction can increase the quality of the benthic reef habitat map.

Table 3: Confusion matrix of 7 benthic habitat classes in the shallow waters of Hurghada city based on ground truthing points and Landsat 8-OLI imagery. Overall Accuracy = 66.7%, % and Overall kappa = 0.611.

C.D.	R. D.								
	C.R.	D.W.	R.B.	S.B.	S.G.	S.W.	Sh.W.		
C. R.	61.33	0.00	0.67	22	18.67	4.00	1.33		
D. W.	0.00	86.67	0.00	1.33	0.00	0.00	13.33		
R.B.	0.67	0.67	69.33	1.33	1.33	0.00	0.00		
S.B.	5.33	0.00	2.67	8.67	0.67	1.33	0.00		
S.G.	8	0.00	26	20.67	63.33	2	0.00		
S.W.	12	0.00	1.33	20.67	16.00	92	0.00		
Sh. W.	12.67	12.67	0.00	25.33	0.00	0.67	85.33		

C.D.: classification data, R. D.: reference data, C.R; Coral Reef, D.W.; Deep Water, R.B.; Rock bottom, S.B.; Sand bottom, S.G.; Seagrass S.W.; Sea weed, Sh.W.; Shallow Water.



Figure 3: Shows different categories used for classification: dense sea weeds; patch coral; live coral; dense seagrass; deep and shallow water; rocky bottom and sandy bottom. Color coding represents habitat classes as described on Table (2).



Figure 4: Benthic habitat classification scheme of Hurghada City. C.R; Coral Reef, D.W.; Deep Water, R.; Rock, S.; Sand, S.G.; Seagrass S.W.; Sea weed, Sh.W.; Shallow Water.

4. DISCUSSION

Remote sensing provides an important, complementary approach to in situ fieldwork for monitoring benthic habitats in shallow water environments. The study revealed that the near shore features as sand, macroalgae, seagrass and coral reef were mostly abundant. Each feature has different spectral characteristics and be separable as homogenous pixels. In reality, there is a significant amount of intermixing between these features. The complex benthic combinations of a mixed sandy/algal cover, mixed rocky/algal and dead coral/turf algae cover areas and error in depth estimation can also have a considerable impact on the classification results. The details deriving ecological and biological information for each field data point would increase the number of elements separable by a classification scheme.

The primary goals of this research were to assess the utility of the Landsat 8-OLI data and to determine the effects of water column and atmospheric correction on image processing techniques, to improve the benthic habitat classification accuracy. The present study showed the capability of increasing the number of benthic habitat classes due to the water column and atmospheric correction. So, the capabilities of the Landsat 8-OLI sensor were increased for benthic habitat mapping. Hyperspectral studies that may aid in discriminating between the mortality states of live and algae covered coral skeletons [47] can be used to develop baseline spectra to help minimize spectral confusion in satellite imagery.

The unsupervised classification gives more classes after masking, water column and atmospheric correction processes. The present investigation indicated that after completed processes of satellite data especially atmospheric and water column correction, total accuracy of coral reef habitat increased and the overall accuracy of the maximum likehood classification increased from 46% to 66.7% and the Kappa coefficient increased from 65.6% to 66.7%. Depending on the level of

classification, previous studies using coarser resolution satellite data (e.g., Landsat TM) have normally achieved accuracies that have ranged from 37 percent [16] to 73 percent [48], even when compensating for the confounding effects of variable water depths. Mumby et al. [16, 49] found the overall accuracy of 4 classes increased from 55% to 73% for (Landsat TM) and 8 classes increased from 38% to 52% (Landsat TM), 13 classes increased from 21% to 37% (SPOT XS), CASI: increased from 72% to 93%. Mumby et al. [16] found the accuracy ranged between 89% and 81% for coarse and fine levels of habitat discrimination while Roelfsema et al. [43] mentioned the overall accuracy 62% increased from 11% to 82%. Andréfouët et al. [33] found the overall accuracy of the classification of IKONOS 77% for 4-5 classes, 71% for 7-8 classes, 65% in 9-11 classes, and 53% for more than 13 classes, for Landsat: 56% for 5-10 classes. Capolsini et al. [50] found that the accuracy of Landsat ETM: increased from 48 to 81%. Purkis et al. [51, 52] mentioned that the overall accuracy was 53% without water column correction and it became 76% after water column correction for Landsat TM while Nurlidiasari [53] found the total accuracy of the images increased from 67% to 89% when the water column correction was considered. Finally, by implementing appropriate atmospheric and water-column corrections, and image classification techniques, coarser-resolution Landsat 8-OLI data become a great tool for mapping benthic habitats.

Acknowledgements

The first author is supported by a scholarship from the Mission Department, Ministry of Higher Education of the Government of Egypt which is gratefully acknowledged. The authors would like to thank the staff of the Remote Sensing Lab at Institute for Marine Remote Sensing (IMaRS), University of South Florida (USF) for critical assistance in acquisition of remote sensing data.

Conflict of Interest

The authors declare that they have no conflicts of interest.

REFERENCES

1. Shaalan IM. Sustainable tourism development in the Red Sea of Egypt threats and opportunities. Journal of Cleaner Production. 2005 13:83-7.

2. Barrania A. Cost of Degradation of Coral reefs and Fisheries Caused by Tourism Development, Egypt's Red Sea: A case study of Hurghada-Safaga Area. Institute of National Planning. Cairo, Nasr City, Egypt; 2010.

3. Lieske E, Myers RF. Coral Reef Guide Red Sea. London: Harper Collins Publishers Ltd; 2004.

4. Hawkins JP, Roberts C. The growth of coastal tourism in the Red Sea: present and future effects on coral reefs. 1994.

5. Frihy O, Fanos AM, Khafagy A, Aesha KAA. Human impacts on the coastal zone of Hurghada, northern Red Sea, Egypt. Geo-Marine Letters. 1996 16:324-9.

6. Green EP, Edwards AJ. Remote Sensing Handbook for Tropical Coastal Management. 2000.

7. Nakaoka M. Distribution of sea grasses over time using anacdotal information of Japanese Fisherman. Japan: Chiba University; 2004.

8. Grabowski J, Bachman M, Demarest CK, Eayrs S, Harris BP, Malkoski V, et al. Assessing the Vulnerability of Marine Benthos to Fishing Gear Impacts. Reviews in Fisheries Science & Aquaculture. 2014 22:142 - 55.

9. Puig P, Canals M, Company JB, Martín J, Amblas D, Lastras G, et al. Ploughing the deep sea floor. Nature. 2012 489:286-9.

10. Olsgard F, Schaanning MT, Widdicombe S, Kendall MA, Austen MC. Effects of bottom trawling on ecosystem functioning. Journal of Experimental Marine Biology and Ecology. 2008 366:123-33.

11. Alberti M, Weeks R, Coe S. Urban Land-Cover Change Analysis in Central Puget Sound. Photogrammetric Engineering and Remote Sensing. 2004 70:1043-52.

12. Xiao J, Shen Y, Ge J-F, Tateishi R, Tang C, Liang Y, et al. Evaluating urban expansion and land use change in Shijiazhuang, China, by using GIS and remote sensing. Landscape and Urban Planning. 2006 75:69-80.

13. Batty M, Howes D. Predicting Temporal Pattern in Urban Development from Remote Imagery. 2001.

14. Herold M, Goldstein NC, Clarke KC. The spatiotemporal form of urban growth: measurement, analysis and modeling. Remote Sensing of Environment. 2003 86:286-302.

15. Khaled MA, Muller-Karger F, Obuid-Allah A, Ahmed M, El-Kafrawy SB. Using Landsat Data to Assess the Status of Coral Reefs Cover along the Red Sea Coast, Egypt. 2019.

16. Mumby P, Green EP, Edwards A, Clark C. Measurement of seagrass standing crop using satellite and digital airborne remote sensing. Marine Ecology Progress Series. 1997 159:51-60.

17. Shalaby A, Tateishi R. Remote sensing and GIS for mapping and monitoring land cover and land-use changes in the Northwestern coastal zone of Egypt. Applied Geography. 2007 27:28-41.

18. Hoffhine Wilson E, Hurd JD, Civco DL, Prisloe MP, Arnold C. Development of a geospatial model to quantify, describe and map urban growth. Remote Sensing of Environment. 2003 86:275-85.

19. Müller D, Zeller M. Land use dynamics in the central highlands of Vietnam: a spatial model combining village survey data with satellite imagery interpretation. Agricultural Economics. 2002 27:333-54.

20. Khaled M. Using Multispectral Satellite Images to Estimate the Environmental State of Coral Reefs at Hurghada Region, Red Sea Coast, Egypt. Egypt: Assiut University; 2013.

21. El-Asmar H, El-Kafrawy SB, Ahmed M, Oubid-Allah AH, Mohamed TA, Khaled MA. Monitoring and Assessing the Coastal Ecosystem at Hurghada, Red Sea Coast, Egypt. Journal of environment and earth science. 2015 5:144-60.

22. Abo-Taleb HA. Mapping the Different Planktonic Groups at One of the Egyptian Bays along Mediterranean Coast. Open Access Journal. 2018 6.

23. Khan MA, Fadlallah YH, Al-Hinai K. Thematic mapping of subtidal coastal habitats in the western Arabian Gulf using Landsat TM dat • Abu Ali Bay, Saudi Arabia. 1992.

24. Michalek JL, Wagner T, Luczkovich J, Stoffle RW. Multispectral change vector analysis for monitoring coastal marine environments. 1992.

25. Ahmad W, Neil D. An evaluation of Landsat Thematic Mapper (TM) digital data for discriminating coral reef zonation: Heron Reef (GBR). 1994.

26. Matsunaga T, Kayanne H. Observation of coral reefs on Ishigaki island, Japan, using Landsat TM images and aerial photographs. 1997.

27. Dobson E, Dustan P. The use of satellite imagery for detection of shifts in coral reef communities. Proceedings of the American Society for Photogrammetry and Remote Sensing Annual Meeting, Washington, D.C; 2000.

28. Peddle DR, LeDrew E, Holden H. Optical correction of scene fractions for estimating regional scale ocean coral abundance in Fiji. IGARSS '96 1996 International Geoscience and Remote Sensing Symposium. 1996 1:427-9 vol.1.

29. LeDrew EF, Holden H, Wulder MA, Derksen C, Newman C. A spatial statistical operator applied to multidate satellite imagery for identification of coral reef stress. Remote Sensing of Environment. 2004 91:271-9.

30. Hamel M, Andréfouët S. Using very high resolution remote sensing for the management of coral reef fisheries: review and perspectives. Marine pollution bulletin. 2010 60 9:1397-405.

31. Klemas V. Remote Sensing Techniques for Studying Coastal Ecosystems: An Overview. 2010.

32. Mumby P, Skirving W, Strong A, Hardy J, LeDrew E, Hochberg E, et al. Remote sensing of coral reefs and their physical environment. Marine pollution bulletin. 2004 48 3-4:219-28.

33. Andréfouët S, Kramer PA, Torres-Pulliza D, Joyce K, Hochberg E, Garza-Perez R, et al. Multi-site evaluation of IKONOS data for classification of tropical coral reef environments. Remote Sensing of Environment. 2003 88:128-43.

34. Mishra D, Narumalani S, Rundquist D, Lawson M. Benthic Habitat Mapping in Tropical Marine Environments Using QuickBird Multispectral Data. Photogrammetric Engineering & Remote Sensing. 2006 72:1037-48.

35. NOAA. What Is a Benthic Habitat Map? <u>http://oceanservicenoaagov/facts/benthichtml</u>. 2004.

36. Brown CJ, Smith SJ, Lawton P, Anderson JT. Benthic habitat mapping: A review of progress towards improved understanding of the spatial ecology of the seafloor using acoustic techniques. Estuarine, Coastal and Shelf Science. 2011 92:502-20.

37. Turner M, Gardner R. Landscape Ecology in Theory and Practice. Springer New York; 2015.

38. Robinson L, Elith J, Hobday A, Pearson R, Kendall B, Possingham H, et al. Pushing the limits in marine species distribution modelling: lessons from the land present challenges and opportunities. Global Ecology and Biogeography. 2011 20:789-802.

39. Hutchinson GE, MacArthur RH. A Theoretical Ecological Model of Size Distributions Among Species of Animals. The American Naturalist. 1959 93:117-25.

40. Cogan CB, Noji T. Marine classification, mapping, and biodiversity analysis. Special Paper - Geological Association of Canada. 2007:129-39.

41. Greene H, Bizzarro JJ, O'Connell V. Construction of Digital Potential Marine Benthic Habitat Maps using a Coded Classification Scheme and its Application *. 2007.

42. Lecours V, Devillers R, Schneider DC, Lucieer V, Brown CJ, Edinger E. Spatial scale and geographic context in benthic habitat mapping: review and future directions. Marine Ecology Progress Series. 2015 535:259-84.

43. Roelfsema C, Phinn S, Dennison W. Spatial distribution of benthic microalgae on coral reefs determined by remote sensing. Coral Reefs. 2002 21:264-74.

44. Pramaditya W, Prama A, Wahyu L. Benthic Habitat Mapping Model and Cross Validation Using Machine-Learning Classification Algorithms. Remote Sensing. 2019.

45. Congalton R. A Quantitative Method to Test for Consistency and Correctness in Photointerpretation. 1983.

46. Mather PM. Review of: @ Introductory Digital Image Processing: A Remote Sensing Perspective By J. R. JENSEN: (Englewood Cliffs, New Jersey: Prentice Hall, 1986) [Pp. 368.] Price £53-45. 1986.

47. Clark CD, Mumby PJ, Chisholm JRM, Jaubert J, Andrefouet S. Spectral discrimination of coral mortality states following a severe bleaching event. International Journal of Remote Sensing. 2000 21:2321-7.

48. Mumby P, Edwards AJ. Mapping marine environments with IKONOS imagery: enhanced spatial resolution can deliver greater thematic accuracy. Remote Sensing of Environment. 2002 82:248-57.

49. Mumby P, Green EP, Edwards A, Clark C. Coral reef habitat mapping: how much detail can remote sensing provide? Marine Biology. 1997 130:193-202.

50. Capolsini P, Andréfouët S, Rion C, Payri C. A comparison of Landsat ETM+, SPOT HRV, Ikonos, ASTER, and airborne MASTER data for coral reef habitat mapping in South Pacific islands. Canadian Journal of Remote Sensing. 2003 29:187 - 200.

51. Purkis SJ, Kenter J, Oikonomou E, Robinson IS. High-resolution ground verification, cluster analysis and optical model of reef substrate coverage on Landsat TM imagery (Red Sea, Egypt). International Journal of Remote Sensing. 2002 23:1677 - 98.

52. Purkis SJ, Pasterkamp R. Integrating in situ reef-top reflectance spectra with Landsat TM imagery to aid shallow-tropical benthic habitat mapping. Coral Reefs. 2003 23:5-20.

53. Nurlidiasari M. The application of QuickBird and multitemporal Landsat TM data for coral reef habitat mapping. Case study: Derawan Island, East

Kalimantan, Indonesia. Netherland: International Institute for Geo-information Science and Earth Observation, Enschede; 2004.

تخريط البيئات القاعية باستخدام بيانات الاستشعار عن بعد بمنطقة الغردقة، ساحل البحر الأحمر، مصر مصطفى خالد'-٢- أحمد عبيدالله فرنك ميلير-كارجير' محمود أحمد سامح الكفراوى' ١- كلية علوم البحار جامعة جنوب فلوريدا – الولايات المتحدة الأمريكية ٢- قسم علوم البحار – الهيئية القومية للاستشعار عن البعد وعلوم الفضاء القاهرة – مصر ٣- قسم علم الحيوان – كلية العلوم جامعة أسيوط – أسيوط - مصر تم تصميم البحث الحالى للتركيز على فائدة بيانات الاستشعار للقمر الصناعي لاندسات (- OLI)المتعددة الأطياف لتحديد وتصنيف وتخريط البيئات القاعية للبحر الأحمر بعد تطبيق التصحيحات الجوية والعمقية لمدينة الغر دقة. تم تطبيق تصحيحات الغلاف الجوى وعمق الماء على الصور، مما بجعلها طربقة فعالة لزبادة القدرة على تخربط البيئات القاعية. تم تحقيق تصحيح العمود المائي للأعماق عن طريق اشتقاق معاملات الامتصاص والتشتت الخلفي لكل نطاق من وحدات البكسل المائية الشفافة للصورة. بعد ذلك تم تطبيق خوارزمية التصنيف الغير مراقب خوارزمية التحليل التكراري الذاتي للبيانات (لإنتاج ٢٢ فئة من البيئات) . تم إجراء التصنيف المراقب باستخدام خوار زمية التعلم الآلية كخوار زمية التشابه القصوى والنقاط المرجعية (البيانات الحقلية) ذلك لإنتاج ٧ فئات من البيئات القاعية على النحو التالي ، الشعاب المرجانية (كثيفة وبقع) ، والأعشاب البحرية (الطحالب الكبيرة) ، والحشائش البحرية (كثيفة وبقع) والمياه العميقة (أكثر من ٢٠ م) والمياه الضحلة (أقل من ٢٠ م) والقاع الرملي (يتكون بشكل أساسى من كربونات الكالسيوم والسيليكا) والقاع الصخرى. أظهرت الأعشاب البحرية (الطحالب الكبيرة) ومناطق المياه العميقة أعلى دقة للبيانات الآلية والمرجعية ، عند مقار نتها بالحشائش البحرية الكثيفة او المختلطة: الحشائش البحرية / الرمل ، والمختلطة: الشعاب المرجانية / الرمال. استنادًا إلى ١٠٥٠ نقطة مرجعية (بيانات حقلية) ، فإن الدقة الإجمالية لتقييم البيئات القاعية تبلغ ٢٦.٧ في المائة ، مع قيمة معامل كابا الاجمالية ٦١١ ...