



ANALYSIS OF BACTERIAL COMMUNITY AND ITS BIODEGRADATION ABILITY OF 17-B ESTRADIOL IN WASTEWATER PLANT IN ASSIUT, UPPER EGYPT

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Wastewater treatment is essential for environmental sanitation in urban environments. Nevertheless, chemical, and organic substances, and microorganisms accumulate in wastewater treatment plants (WWTPs), which may be liberated into the environment via their effluents. This study aimed to analyze the bacterial community at the sewage treatment plant entrance in Assiut, Upper Egypt as well as their degradation ability of 17- β estradiol (E2) as an Isolates were identified by conventional methods and, 16S rRNA gene sequencing. E2 degradation by different strains was assessed using HPLC/UV. Conventional methods and 16S rRNA based amplicon sequencing revealed three bacteria including *Bordetella sputigena* (ASc1), *Pigmentiphaga daeguensis* (ASc3), and *Stenotrophomonas pavanii* (ASc4), and when tested for E2 degradation, they degraded 65.5%, 59%, and 66.4% of E2 (50 mg. L⁻¹) respectively. Other bacterial strains were identified using 16S rRNA based amplicon sequencing. Bacteroidetes, Firmicutes and Proteobacteria were the prominent phyla. However, Actinobacteria, Cyanobacteria, Chloroflexi, Caldithrix, and Chlamydiae were detected in lower proportions.

Keywords: Bacterial community composition; Wastewater treatment plants (WWTPs); 17- β estradiol (E2); Biodegradation

INTRODUCTION

All communities of bacteria which are responsible for the degradation of carbon and the elimination of nutrients from the wastewater are unknown. These nutrients are the core elements of any biological WWTP^{1,2}. Most waterborne microorganisms spread via the fecal-oral route and enter sewage treatment plants through human excretion of feces².

Endocrine-disrupting chemicals (EDCs) involve various kinds of synthetic and natural chemicals which stimulate or inhibit the reproductive effects of the endocrine system in humans and animals. Several important EDCs have been classified into two main classes: natural estrogens like estrone (E1), E2, and estriol (E3) and synthetic estrogens like ethinyl

estradiol (EE2)³. The impacts of these compounds in humans include Infertility, defined cancers, decreased sperm number, changed testosterone levels, intrauterine growth restriction and preterm labor, decreased fetal testosterone, decreased male fertility, urogenital disorders such as cryptorchidism and hypospadias, recurrent miscarriages, testicular cancer, birth weight loss, polycystic ovary syndrome, genital abnormalities, breast and prostate cancer, endometriosis, early puberty, infertility, menstrual disorders, and obesity^{4,5}.

E2 is one of the natural estrogens excreted by humans and cattle, which in very low concentrations (nanogram/litre) has an endocrine effect and causes serious problems in the ecosystem^{6,8}. Women excrete estrogen from their bodies without using external hormonal

drugs^{7,10}. Ying et al. found that women and men excreted the hormone estradiol at a rate of 1.6 µg/day⁸. Therefore, the presence of this hormone in wastewater has become normal and should be treated as a problem to be solved.

Many processes are involved in the elimination of estrogens from wastewater, involving biological, physical and chemical removal^{9,13}. In plants of sewage treatment, the elimination of estrogens by microorganisms under aerobic conditions through conversion reactions and biological metabolism plays an important role, and by which the bacteria use estrogens in biochemical reactions as a source of carbon^{10,11} and^{12,17}.

The microbial communities in the sewage treatment plants were determined by traditional methods (culture-dependent methods) such as observation under an optical microscope^{13,19}, or by performing various biochemical tests on isolated strains^{14,21}. Current methods for the detection of microorganisms (culture-independent methods), like Next-generation sequencing, were used for identification purposes by extraction of genomic DNA from wastewater samples and 16S rRNA gene sequencing¹⁵.

In this work, bacterial populations in wastewater were identified using culture-dependent or culture-independent techniques and the ability of each isolate to degrade E2 alone was evaluated.

MATERIAL AND METHODS

Materials

Estradiol (E2; >98% purity) was obtained from Sigma-Aldrich. (USA). Acetonitrile and methanol was bought from TEDIA that were pure enough for high-performance liquid chromatography (HPLC). The E2 was made as a stock solution with a high concentration of methanol (1000 mgL⁻¹).

The medium of mineral salt (MSM)¹⁶ consisted of 0.50 g KH₂PO₄, 1.50 g (NH₄)₂SO₄, 1.91 g K₂HPO₄·3H₂O, 0.20 g MgSO₄·7H₂O, 0.5 g NaCl, and 1 mL trace element solution in 1 L ultra-pure water. The trace element solution comprised the following amounts of each element: 0.10 g/L CoCl₂·6H₂O, 0.015 g/L CuSO₄·5H₂O, 0.50 g/L ZnCl₂, 0.425 g/L MnCl₂·4H₂O, 0.01 g/L NiCl₂·6H₂O, and 0.01 g/L Na₂MoO₄·2H₂O.

The estrogen mineral salt medium (EMM) contained MSM augmented with E2 at the test doses and had a pH of 7.0 ± 0.2. E2 was placed into sterilized flasks with the required concentrations, and the methanol was permitted to volatilize. Following that, the MSM was appended to generate the liquid EMM. For the solid EMM plates, the melted MSM was combined with E2, the medium was put onto the plates, and the methanol was then volatilized with a stream of sterile air. Agar at a concentration of 18.0 g/L was added to the liquid medium to create solid media plates. Mannitol salt agar (Oxoid, UK) was used for the isolation of *Staphylococci*¹⁷. While MacConkey agar (Oxoid, UK) was used for isolation of Gram negative bacteria¹⁸ and Cetrimide agar (Becton Dickinson, U.S.A) was used for isolation of *pseudomonas aeruginosa*¹⁹.

METHODS

Enrichment procedure for bacterial isolation

The enrichment procedure was done in a 250-mL Erlenmeyer flask involving 100 mg. - L⁻¹ E2 in 100 mL of liquid MSM²⁰. One hundred water specimens were collected over six months (from December 2020 to May 2021), from the main sewage water drain (Sewage treatment plant, Arab El- Madabegh, Assiut, Upper Egypt). Wastewater specimens were collected from a depth of 0.5 m. Ten samples (each one is 0.5 liter) were obtained from different positions at one time were collected in sterile, nonreactive borosilicate glass bottle to form 5 liter pooled together, and prefiltered through cheesecloth to remove debris, then filtered on a sterile membrane filter with a pore size of 0.22 µm. Filter papers were cultured in a mineral salt medium (MSM), and the mixture was grown in an aerobic incubator shaking at 150 r.min⁻¹ and at 30 °C. Five percent (v/v) of this combination was put into new liquid MSM at 100 mg. - L⁻¹ E2, every 2 days, and this procedure was carried out three times. The bacterial isolates were cultured for 48 hours at 30 °C after being seeded into independent EMM plates with 100 mg.- L⁻¹. Repetitive streaking was used to choose morphologically different colonies for purification before a single purified colony was

produced and transferred to maintenance slant cultures²⁰.

Harvesting of bacterial cells. To create cell suspension, every purified unique colony was increased by culturing it in Luria Bertani (LB) medium on a rotary shaker at 150 r.min⁻¹ and at 30 °C. The cell suspension was centrifuged at 8000 rpm min⁻¹ for 4 min, after which the pellet containing bacteria was gathered and twice rinsed with phosphate buffer saline (PBS). It was then put back together in sterile MSM and given an optical density of 1 at 600 nm²⁰.

Identification of isolated strains

Three bacterial isolates named as ASc1, ASc3 and ASc4 were isolated from all samples, and subjected to Gram stain, culture on Cetrimide agar, Mannitol salt agar, and MacConkey agar and biochemical tests including citrate utilization, catalase, DNase, oxidase and indole production for characterization and identification²¹. The genomic DNA was extracted by utilizing Gene JET Genomic DNA Purification Kit #K0721 (thermo scientific, Lithuania). The following universal primer pairs^{22,23} (sequence 5'-3') 27 forward (F) 5'-AGA GTT TGA TCC TGG CTC AC-3' and 1492 reverse (R) 5'-GGT TAC CTT GTT ACG ACT T-3' were utilized for amplification of 16S rRNA gene by PCR. The following profile was used for the amplification: initial denaturation for 5 min at 95°C, followed by 30 cycles of denaturation for 30 seconds at 95°C, annealing for 30 seconds at 57°C, extension for 1 min at 72°C, and final extension for 10 min at 72°C²⁴. Amplification products were sequenced at GATC Company utilizing ABI 3730xl DNA sequencer utilizing forward and reverse primers. 16S rRNA gene sequences from isolates were submitted and aligned to the gene sequences available in the NCBI database utilizing the BLAST server (<http://blast.ncbi.nlm.nih.gov/>). The tree built from aligned 16S sequences of stems that closely match the searched database sequences²⁵.

Screening the ability of ASc1, ASc3 and ASc4 strains for E2 biodegradation

The bacterial capacity for biodegradation of E2 was evaluated by growing it aerobically in an EMM medium. The degradation system

consisted of (50 mg.- L⁻¹ E2) a 5% (v/v) cell suspension in 20 ml of EMM in 50 ml Erlenmeyer flasks. The combination was cultivated on a rotary shaker for 48 hrs at pH 3 and 150 rpm.- (7.00). Residual E2 in bacterial solutions was identified by HPLC/UV every 12 hours as follows; 20 mL of methanol was appended to flasks containing bacterial cultures to lyse the rest of E2, and the combination was sonicated for 30 min and then filtered via 0.22 µm polytetrafluoroethylene membrane filters. These settings were utilized for HPLC/UV identification; column C₁₈, a motile stage of acetonitrile and water (70/30 v/v) at a flow rate of 1 mL/min and an injection volume of 20 µL²⁶.

The degradation efficiency (%) of E2 in batch culture was calculated as follows:

$$\text{Degradation efficiency (\%)} = \frac{(C_{\text{Initial}} - C_{\text{Residual}})}{C_{\text{Initial}}} \times 100\%.$$

Where C_{Initial} and C_{Residual} represent the primary and last concentrations of estradiol in culture solution.

Microbial community characterization

Genomic DNA extraction and 16S rRNA based amplicon sequencing

The samples of wastewater were pre-filtered with a cheesecloth before extraction to remove debris and then filtered through a sterile membrane filter (0.2 µm). Total genomic DNA was isolated from membrane filters by utilizing the DNeasy PowerSoil Kit Cat No. 12888-100 (Qiagen, USA) as per the manufacturer's instructions.

PCR was used to amplify up the isolated DNA utilizing primers specific for the V1–V3 regions. Primer-attached Illumina adaptor (underlined),

Forward	Primer	5'
AATGATACGGCGACCACCGAGATCTAC		
AC	ATCGTACGTAT	
GGTAATTCAATTACCGCGGCTGCTGG		
Reverse	Primer	5'
CAAGCAGAAGACGGCATACGAGAT		
AACTCTC		
GAGTCAGTCAGCCGAGTTTGAT		
CMTGGCTCAG ²⁷ .		

An initial denaturation for 2 min at 95°C, followed by 22 cycles of denaturation for 20 seconds at 95°C, annealing for 15 seconds at 55°C, extension for 1 min at

72°C, and the last extension for 10 min at 72°C, formed the PCR amplification protocol²⁸. Utilizing Agencourt XP Ampure Beads (Agencourt Bioscience Corp., Beverly, MA), amplification materials were purified and sequenced with the Illumina MiSeq sequencing system at IGA Technology Service (Udine, Italy).

Sequence processing and analysis

The Illumina MiSeq platform generated data in fastq file format for forward and reverse paired-end reads without barcode sequences. The Quantitative Insights Into Microbial Ecology (QIIME2)^{29,30} bioinformatics pipeline was used to analyze the 16S rRNA gene readouts. Raw sequence preparation depended on ribosomal sequence variants (RSVs), which are sequences with differences of more than one nucleotide. Quality control was performed using the DADA2 plugin of the Qiime2 platform³¹. In summary, demultiplexed coupled reads from the research specimens and the negative controls were used as input for filtering and clipping reads with an absolute peak of dual predicted faults per read (maxEE= 2) and a diminished length for reverse and forward reads of 210 and 270, respectively. At the final stage of the readings, discard low-quality bases, and demand an average Phred quality of at least 25. Then the filtered reads were dereplicated into distinguishable sequences, and the DADA2 inference algorithm used the 1000000 parameters of the DADA2 error model to denoise each sequence. When merging denoised reverse and forward sequences, a minimal overlapping of 20 base pairs and zero mismatches were used. Potential chimeric sequences were eliminated after ribosomal sequence variants (RSVs) were developed to create high-resolution RSVs. Representative of RSVs Sequences differed by >3% from each other. The resulting representative sequences were used for subsequent taxonomic assignments and diversity metrics.

Representative sequences of every RSV were utilized for taxonomy categorization versus the SILVA SSU Ref NR dataset v.132³² utilizing trained RDP's naive Bayesian classifier at 99% sequence similarity^{33,37}. Instead, ESPRIT-Tree software used taxon-independent clustering^{34,34} to group sequences into operational taxonomic units (OTUs) by calculating the k-mer distance across sequences at similar values ranging from 99% (species/strain stage) to 80% (phylum stage).

RESULTS AND DISCUSSION

Results

Identification of isolated strains

Three strains of bacteria; ASc1, ASc3 and ASc4 were isolated from wastewater. All three isolates were grown on MacConkey agar and pure gram-negative bacilli. Analysis of the biochemical properties of the isolated strains showed that; The ASc4 isolate lacked the ability to grow on Cetrimide agar, Mannitol salt agar and was negative for oxidase, indole, ornithine decarboxylase and citrate, and positive for lactose fermentation and catalase. Despite ASc1, ASc3 isolates lacked the ability to grow on Cetrimide agar, Mannitol salt agar and were negative for indole, lactose fermentation, ornithine decarboxylase and citrate, yet were - positive for catalase and oxidase.

16S rRNA gene sequences from isolates were submitted and aligned to gene sequences available in the NCBI database utilising the BLAST server. A BLAST analysis revealed that the strains ASc1, ASc3 and ASc4 were 95.43%, 87.96% and 99.43% identical to *Bordetella sputigena*, *Pigmentiphaga daeguensis* and *Stenotrophomonas pavaini* respectively. **Fig.s (1-3)** show a phylogenetic tree of strains ASc1, ASc3 and ASc4 and related strains.

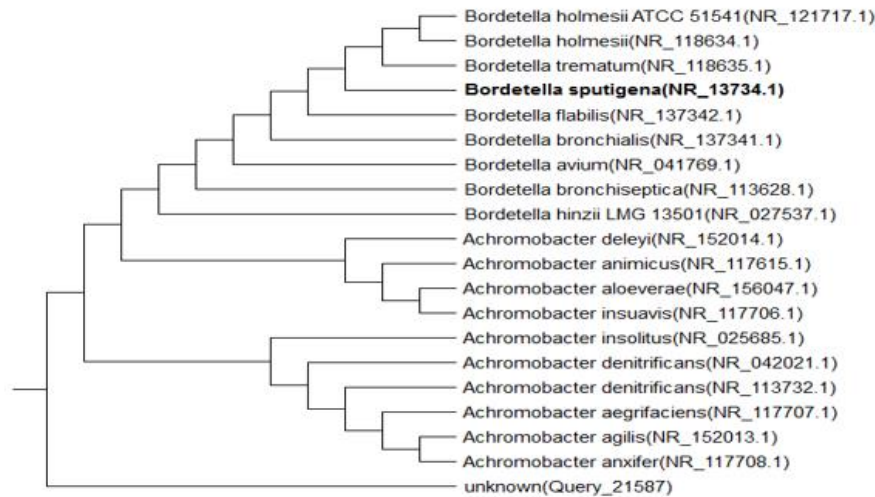


Fig. 1: Phylogenetic tree of *Bordetella sputigena* and closely linked species.

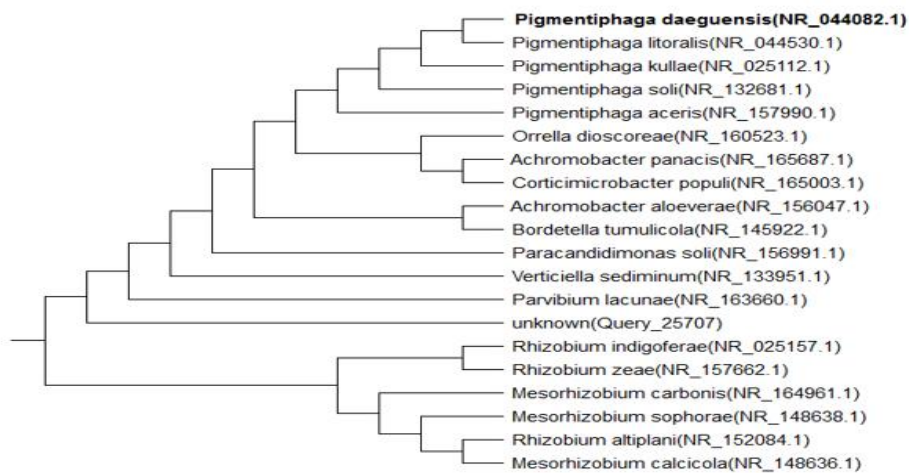


Fig. 2: Phylogenetic tree of *Pigmentiphaga daeguensis* and closely linked species.

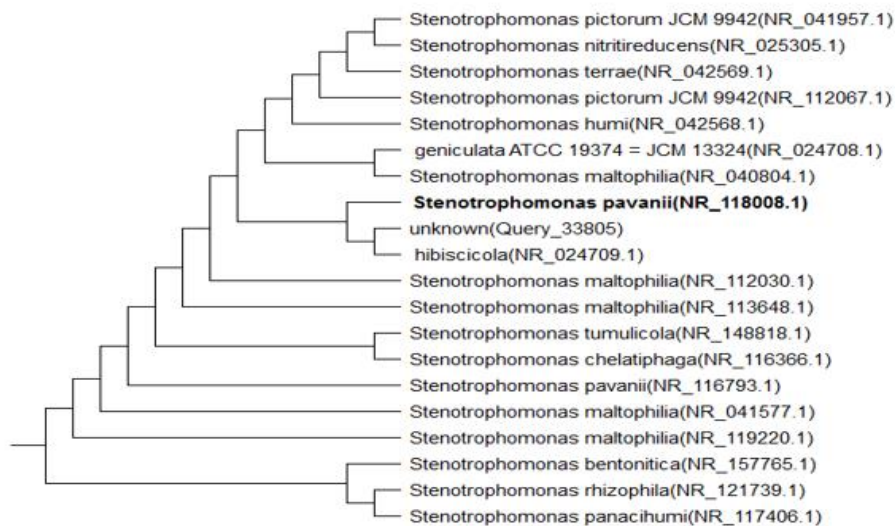


Fig. 3: Phylogenetic tree of *Stenotrophomonas pavanii* and closely linked species.

Degradation of E2 by ASc1, ASc3 and ASc4 strains.

The rate of E2 degradation by strains ASc1, ASc3 and ASc4 strains was evaluated separately for each strain. ASc1, ASc3 and ASc4 strains showed E2 biodegradation efficiencies of 65.5%, 59% and 66.4%, respectively, with an initial E2 concentration of 50 mg.- L⁻¹ after 48 hrs, as shown in Fig.(4).

Bacterial diversity analysis at WWTP

Bacterial diversity was analyzed depending on 16S rRNA gene sequence analysis, leading to a total of 11, 979, 250 reads throughout the ten specimens. After quality filtering, removal of reads $420 \leq$ and ≥ 470 , removal of duplicate sequences and removal of chimeric sequences, and there were 1, 669, 510 reads as representative sequences that were unique sequences in our dataset, as shown in Table (1).

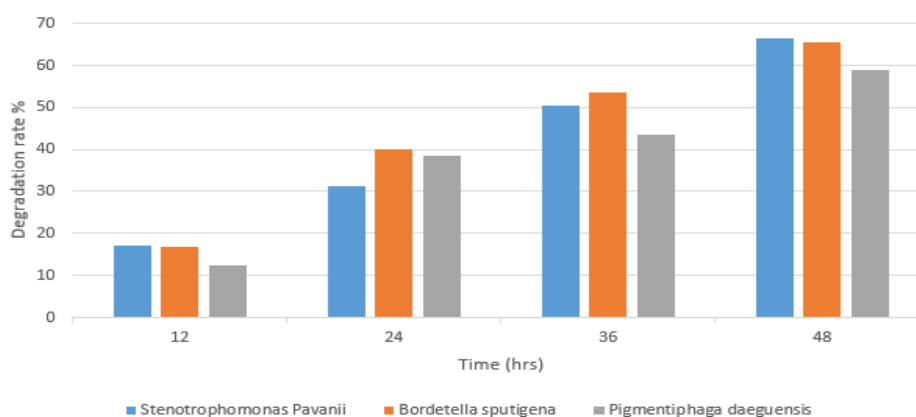


Fig. 4: Degradation of E2 by *Bordetella sputigena*, *Pigmentiphaga daeguensis* and *Stenotrophomonas pavaini* in MSM with E2 at concentration of (50 mg.- L⁻¹) during 48 hours of incubation.

Table 1: Sequence analysis for the generated reads and the remaining reads after preprocessing and filtration.

Sample Code	Number of reads	No. of reads after quality and length filtering	No. of representative sequences	No. of reads after chimeras elimination	Percentage of chimeras sequences
ASu1	784,721	731,502	133,342	125,341	6 %
ASu2	690,162	643,318	109,339	104,474	4.2 %
ASu3	820,266	769,499	131,241	122,447	6.7 %
ASu4	339,702	316,396	60,394	58,038	3.9 %
ASu5	1,577,319	1,428,655	257,121	238,351	7.3 %
ASu6	1,753,896	1,607,497	258,996	242,420	6.4 %
ASu7	440,319	410,107	61,493	58,787	4.4 %
ASu8	1,406,148	1,282,586	179,537	168,944	5.9 %
ASu9	1,915,402	1,752,631	262,867	242,889	7.6 %
ASu10	2,251,315	2,049,063	327,816	307,819	6.1 %

Bacterial community structure at WWTP

At the phylum, class, order, family, and genus stages, analyses of the relative abundances of bacterial groupings were conducted. Ten samples yielded 28 bacterial phyla, 53 classes, 104 orders, 226 families, and 559 genera. **Fig. (5)** displays the bacterial community's architecture and relative abundance at the phylum level. As indicated in **Fig. (5)**, the general sewage sequences have been classified into three dominant phyla, namely Bacteroidetes, Firmicutes and Proteobacteria. Samples ASu6, ASu7, ASu8 and ASu10 had the highest frequency of Bacteroidetes, representing (49%), (50%), (55%) and (56%) respectively, while samples ASu1, ASu2, ASu3 and ASu4 had the highest frequency of Firmicutes accounting for (56%), (49%), (45%) and (51%) respectively. Proteobacteria represented approximately (27%), (26%), (25%), (24%) and (18%) in

ASu7, ASu8, ASu6, ASu9 and ASu10 samples, respectively.

At the class level, the dominant classes among Bacteroidetes were Bacteroidia, Flavobacteria and Sphingobacteria. Bacteroidia class represents (49.7%) of all sequences, most Bacteroidetes sequences are closely related to Genera Prevotella (39.8%), Bacteroides (6.3%), Parabacteroides (2.1%), Flavobacterium (2.8%) and Sphingobacterium (0.61%) as shown in **Fig. (6)**. Regarding the Firmicutes phylum, three main classes have been found within this phylum, Clostridia, Bacilli and Erysipelotrichi. The Clostridia class represents (50.1%) of all sequences that were considered the highest dominant class, while sequences assigned to Bacilli in the lowest dominant class averaged (5.1%). At the genus level, most Firmicutes sequences were related to Clostridium (13.46%), Blautia (5.3%) and Lactobacillus (3.9%), as shown in **Fig. (7)**.

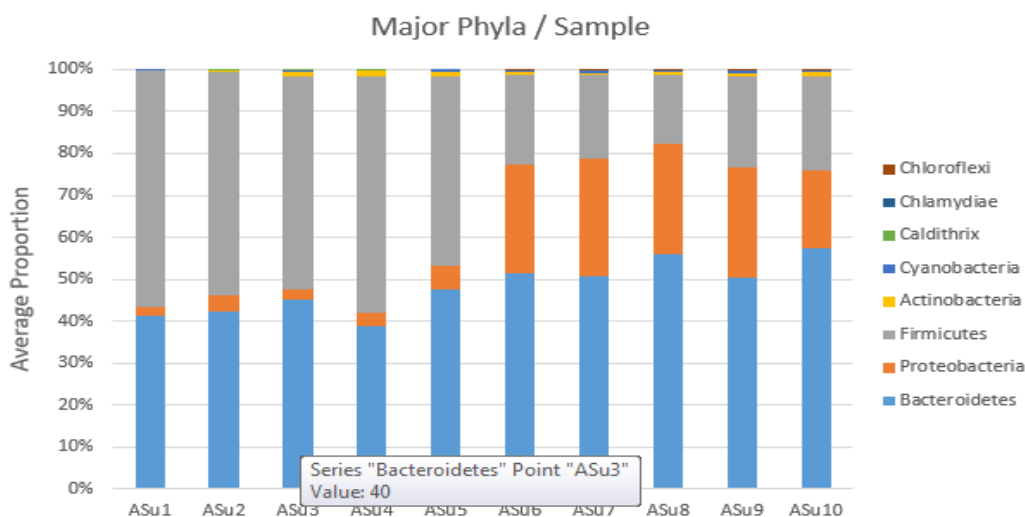


Fig. 5: Distribution of most predominant phyla in wastewater samples.

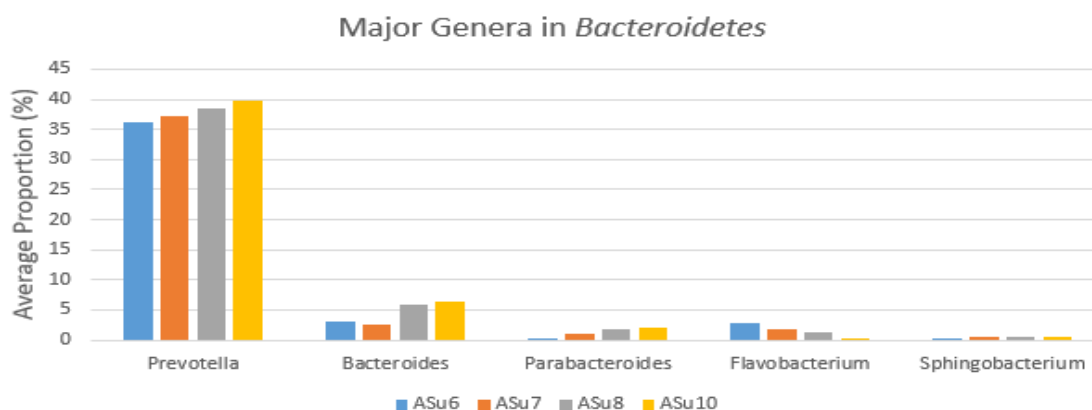


Fig. 6 : Distribution of most abundant genera in Bacteroidetes phylum.

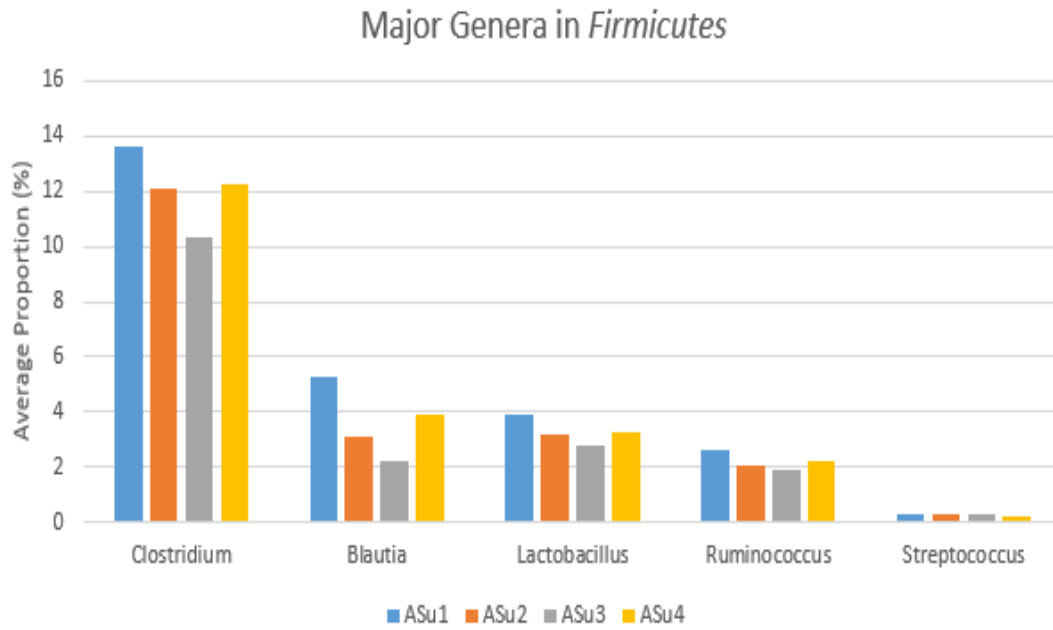


Fig. 7: Distribution of most abundant genera in Firmicutes phylum.

Regarding the Proteobacteria phylum, most of the bacterial sequences belong to the Gammaproteobacteria, Epsilonproteobacteria, Betaproteobacteria, Deltaproteobacteria, and Alphaproteobacteria classes, as shown in **Fig. (8)**. At the species level, *Arcobacter Cryaerophilus*, *Arcobacter Defluvii*, *Escherichia Albertii*, *Succinivibrio Dextrinosolvens* and *Acinetobacter Johnsonii*

were the predominant species under the *Proteobacteria* phylum as shown in **Fig. (9)**.

Additionally, five phyla; Actinobacteria, Cyanobacteria, Chloroflexi, Caldithrix and Chlamydiae were detected in low proportions in all samples (0.2% to 1.1%), (0.1% to 0.5 %), (0.005% to 0.3%), (0.001% to 0.1%) and (0.001% to 0.01%) respectively.

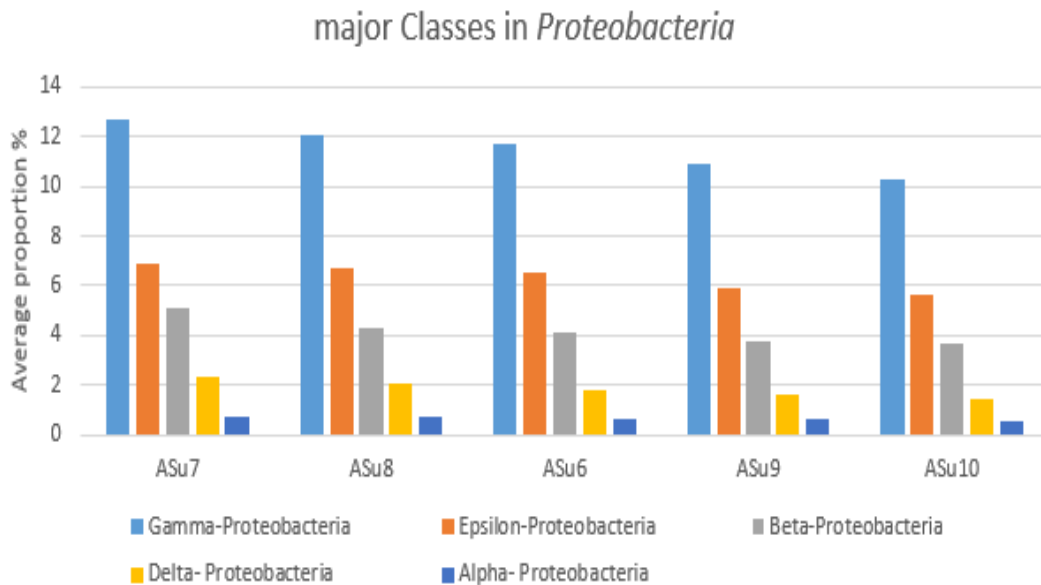


Fig. 8 : Distribution of classes in Proteobacteria phylum.

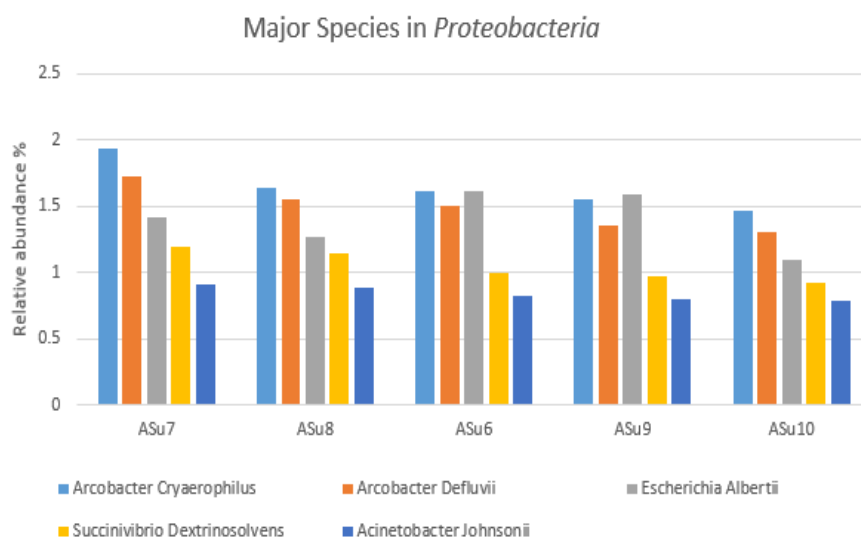


Fig. 9 : Relative abundance of major Species in Phylum Proteobacteria.

Unclassified sequences

At various taxonomic stages, there were sequences which could not be allocated to the reference sequence based on comparison with the SILVA reference database at 0.97 and 0.99% similarity. The number of unclassified sequences increased with the focus on more informative taxonomic levels such as species or genus stages.

At the phylum level, unassigned sequences accounted for 0.78% of all reads, while at the class level, they accounted for 1.60 %. The number of unclassified sequences increased steadily with the taxonomic classification by 1.94 %, 5.36 % and 7.45 % at the order, family and genus stages, respectively. At the species level, unclassified sequences accounted for 39.1 % of all sequences.

Discussion

Treatment of wastewater is one of the most important biotechnological approaches applied to treat domestic and industrial wastewater throughout the world. WWTPs are intended to lower overall bacterial loads in addition to sediments and nutrients^{36,37}. Many toxic compounds (chemicals and organics) and bacteria that have been collected in sewage treatment facilities may be discharged into the environment through the effluent from WWTPs. As a result of improving microbial populations in wastewater treatment systems, numerous hazardous chemicals and diseases are effectively removed^{38,43}.

In developing countries, knowledge of the overall microbial communities present in WWTPS is very limited. The investigation of the function and structure of the bacterial community of these populations has become a critical component of the biotechnological procedures employed in sewage treatment plants³⁹.

To quantify the bacterial diversity in wastewater specimens, we employed both culture-independent and culture-dependent techniques in this research. In terms of culture-dependent approaches, three strains; *Bordetella sputigena*, *Pigmentiphaga daeguensis* and *Stenotrophomonas pavaini* were identified in a culture medium using E2 as the sole source of carbon. These isolates showed E2 biodegradation efficiencies of 65.5%, 59% and 66.4%, respectively, at an initial E2 concentration of 50 mg. L⁻¹ after 48 hrs. This finding is similar to that of Zhou et al., who identified three bacterial isolates (*Isoptricola* spp. (EsD8), *Nubsella* spp. (EsD18), and *Rhodococcus* spp. (EsD20)) from wastewater samples that could decompose E2⁴⁰ and by Fernandez et al., who isolated five bacteria from sediment enrichment cultures containing E2 as the only carbon source. Three of these were identified as *Bacillus flexus*, *Virgibacillus halotolerans*, and *Bacillus licheniformis*⁴¹. A prior research by Mainetti et al. noted that 26 isolates (including *Rhodococcus* spp. *Pseudomonas* spp., and *Alcaligenes* spp) were isolated from plant effluents in Switzerland by

growing them in a liquid medium containing E2 as the sole source of carbon⁴². In China, the species *Orchobacterum* (strain FJ1) was isolated from WWTP and demonstrated its ability to remove $98 \pm 1\%$ of the initial E2 concentration after 10 days⁴³.

Previous Egyptian studies showed diverse results. For example, khatab *et al.*, who performed their study on sewage water at Saft El-Henna (Northern Egypt), demonstrated six bacterial strains that degraded E2 involving *Klebsiella* sp., *Bacillus* sp., *Enterobacter* sp., *Enterobacter* sp II., *Aeromonas veronii*, and *Aeromonas punctata*⁴⁴. In another study, Sewage water from the Industrial Water Works in the Tenth of Ramadan City (El-Sharkia Governorate) tested positive for *Pseudomonas* sp⁴⁵. Also, *Bacillus subtilis* was found in association with *Lactobacillus plantarum*, *Pediococcus acidilactici*, and *Pediococcus pentosaceus* in Quhafa Wastewater Treatment Plant (QWWTP), Fayuom City⁴⁶.

In the culture-independent methods, 28 bacterial phyla were found in ten specimens. Firmicutes, Bacteroidetes and Proteobacteria were the most common phyla. Bacteroidetes was the most dominant phylum in ASu6, ASu7, ASu8 and ASu10 samples representing (49%), (50%), (55%) and (56%) of the classified sequences, respectively. Firmicutes was the mostly second abundant phylum, comprising (56%), (49%), (45%) and (51%) of the classified sequences in ASu1, ASu2, ASu3 and ASu4 samples, respectively, while Proteobacteria was the mostly third abundant phylum representing about (27%), (26%), (25%), (24%) and (18%) in samples ASu7, ASu8, ASu6, ASu9 and ASu10 respectively. This is in agreement with the observations mentioned by Silva-Bedoya *et al.*, who reported that the commonest microorganisms in WWTPs were Firmicutes, Proteobacteria and Bacteroidetes phyla⁴⁷. Previous research by Xu *et al.* reported a total of 36 phyla 8 specimens. Bacteroidetes, Actinobacteria, Proteobacteria, and Chloroflexi were the most predominant phyla, which accounted for (19.3 – 37.3%), (1.5 – 13.8%) (26.7 – 48.9%), and (2.9 – 17.1%) of the total sequences in every specimen, respectively⁴⁸. Also, Shu *et al.* noted that Bacteroidetes, Firmicutes, Proteobacteria, and Chloroflexi were the most common phyla in six specimens from WWTPs⁴⁹.

However, on the other hand to the current work, some investigators reported that Proteobacteria was the phylum that was detected in three WWTPs to be the most frequently occurring, accounting for (11-85%), followed by Firmicutes (0.8-82.9%), Actinobacteria (3.9-16.1%), Bacteroidetes (0.1-5%) and Chloroflexi (0.1-2.5%) in South Africa⁵⁰. In Germany, Firmicutes was the phylum with the highest average abundance among sequences found in samples from WWTPs with an average ($52.2 \pm 4.4\%$) followed by Proteobacteria ($37.8 \pm 4.7\%$), Bacteroidetes ($4.9 \pm 1.9\%$) and Actinobacteria ($2.2 \pm 0.2\%$)⁵¹. Proteobacteria was also the most common phylum in China, representing (59.63%) of total sequences discovered in WWTPs, followed by Bacteroidetes, Acidobacteria, Firmicutes, and Nitrospirae⁵². In addition, water samples taken from Lake Qarun in Egypt included five different bacterial phyla in varying proportions: Firmicutes (53%), Proteobacteria (33%), Bacteroidetes (7%), Actinobacteria (5%) and Thermi (1%)⁵³.

In this study, from winter (43% relative abundance) to spring (56% relative abundance), the relative abundance of Bacteroidetes rose steadily. A rise from winter's 5% to spring's 27% relative abundance of Proteobacteria was also observed. While the relative abundance of Firmicutes grew dramatically from 22% in the spring to 56% in the winter. The results of this study are in line with those of a previous one⁵⁴, which found that seasonal shifts had an effect on the composition of bacterial communities. The results of the wastewater analysis also showed that the total bacterial counts were higher in El Mehala Elkobra throughout the summer. Isolated microorganisms include *Aeromonas sobria*, *Staphylococcus aureus*, *Salmonella* sp., *Proteus mirabilis*, *Serratia plymuthica*, and *Pantoea agglomerans*⁵⁵.

In this study, the dominant classes at ten wastewater specimens were Bacteroidia, Flavobacteria and Sphingobacteria, Clostridia, Bacilli, Erysipelotrichi, Gammaproteobacteria, Epsilonproteobacteria, Betaproteobacteria, Deltaproteobacteria and Alphaproteobacteria. These findings are consistent with a prior study⁵⁰ stating that the dominant classes at the 6 wastewater specimens were Actinobacteria, Alphaproteobacteria, Bacilli,

Gammaproteobacteria, Clostridia, Deltaproteobacteria, Bacteroidia, Flavobacteria, and Betaproteobacteria. In addition, Tian et al. (2017) from China claimed that the key efficient classes were Alphaproteobacteria, Anaerolineae, Gammaproteobacteria, Betaproteobacteria, and Flavobacteria⁵⁶. In contrast, Oyedara et al. (2022) in Mexico noted the distribution of classes Epsilonproteobacteria, Bacteroidia, Fusobacteria, Erysipelotrichi, Planctomycetacia, Sphingobacteria, and Nitrospirae in wastewater samples⁵⁷. Observed microbial diversity between this study and other studies may be attributed to geographical differences and different climate conditions.

Conclusion

Three different gram-negative bacterial strains (ASc1, ASc3 and ASc4) isolated from wastewater were identified as E2-degrading bacteria by a culture-dependent method. Analysis of the 16S rRNA gene sequences of the extracted strains identified them as belonging to the genus *Bordetella sputigena*, *Pigmentiphaga daeguensis* and *Stenotrophomonas pavaini*, respectively. ASc1, ASc3 and ASc4 strains exhibited E2 biodegradation efficiency of 65.5%, 59% and 66.4%, respectively, at an initial E2 concentration of 50 mg. L⁻¹ after 48 hrs. according to the culture-independent method, 28 bacterial phyla, 53 classes, 104 orders, 226 families and 559 genera were detected from ten samples. The most common phyla in all sequences were Bacteroidetes, Firmicutes and Proteobacteria. Bacteroidia was the dominant class among Bacteroidetes and Clostridia was the major class among Firmicutes, while most bacterial sequences in the phylum Proteobacteria were the classes of Gammaproteobacteria, Epsilonproteobacteria, Betaproteobacteria, Deltaproteobacteria and Alphaproteobacteria.

Recommendations

Because E2-degrading strains have demonstrated their ability to reduce the potential environmental risks posed by E2, the removal and biodegradation of ecosystem pollutants should be investigated further.

· More knowledge of the entire microbial communities found in sewage water will aid

researchers in improving the efficacy of biodegradation of environmental contaminants.

Abbreviations

ASc1	<i>Bordetella sputigena</i>
ASc3	<i>Pigmentiphaga daeguensis</i>
ASc4	<i>Stenotrophomonas pavaini</i>
C _{Initial}	the initial concentrations of estradiol in culture solution
C _{Residual}	the final concentrations of estradiol in culture solution.
E1	estrone
E2	17 β-estradiol
E3	Estriol
EDCs	Endocrine disrupting chemicals
EE2	ethinyl estradiol
EMM	estrogen mineral salts medium
LB	Luria Bertani
MSM	mineral salt medium
OTUs	operational taxonomic units
PBS	phosphate-buffered saline
QIIME2	The Quantitative Insights Into Microbial Ecology
RSVs	ribosomal sequence variants
WWTPs	wastewater treatment plants

REFERENCES

1. R. Wirth, B. Pap, T. Bőjti, P. Shetty, G. Lakatos, Z. Bagi, *et al.*, "Chlorella vulgaris and its phycosphere in wastewater: Microalgae-bacteria interactions during nutrient removal", *Front Bioen Biotechnol*, 8, 557572 (2020).
2. F. Y. Ramírez-Castillo, A. Loera-Muro, M. Jacques, P. Garneau, F. J. Avelar-González, J. Harel, *et al.*, "Waterborne pathogens: detection methods and challenges", *Pathogens*, 4(2), 307-34 (2015).
3. C. Zhang, Y. Li, C. Wang, L. Niu and W. Cai, "Occurrence of endocrine disrupting compounds in aqueous environment and their bacterial degradation", *Crit Rev Environ Sci Technol*, 46(1), 1-59 (2016).
4. S. Leslie, T. Soon-Sutton and MA. Khan, "Male infertility", *In: StatPearls (Interent), Treasure Island (FL)*, (2023).
5. D. Balabanič, M. Rupnik and A. K. Klemenčič, "Negative impact of

- endocrine-disrupting compounds on human reproductive health", *Reprod. Fertil. Dev.*, 23(3), 403-16 (2011).
6. B. Almazrouei, D. Islayem, F. Alskafi, M. K. Catacutan, R. Amna, S. Nasrat, *et al.*, "Steroid hormones in wastewater: Sources, treatments, environmental risks, and regulations", *Emerg Contam.*, 9(2), 100210 (2023).
 7. M. Adeel, X. Song, Y. Wang, D. Francis and Y. Yang, "Environmental impact of estrogens on human, animal and plant life", A critical review. *Environ Int.*, 99, 107-119 (2017).
 8. G. G. Ying, R. S. Kookana, Y. J. Ru, "Occurrence and fate of hormone steroids in the environment", *Environ Int.*, 28(6), 545-551 (2002).
 9. D. Guerrero-Gualan, E. Valdez-Castillo, T. Crisanto-Perrazo and T. Toulkeridis, "Methods of Removal of Hormones in Wastewater", *Water*, 15(2), 353 (2023).
 10. Y. L. Chen, H. Y. Fu, T. H. Lee, C. J. Shih, L. Huang, Y. S. Wang, *et al.*, "Estrogen degraders and estrogen degradation pathway identified in an activated sludge", *Appl Environ Microbiol.*, 84(10), e00001-18 (2018).
 11. Y. Koh, T. Chiu, A. Boobis, E. Cartmell, M. Scrimshaw and J. Lester, "Treatment and removal strategies for estrogens from wastewater", *Environ Technol.*, 29(3), 245-267 (2008).
 12. J. Guo, S. Qiu, L. Dai, L. Zhang, L. Meng, M. Liu, *et al.*, "The occurrence and removal of steroid estrogens in a full-scale anaerobic/anoxic/aerobic-membrane bioreactor process and the implication of the bacterial community dynamics", *J Environ Chem Eng.*, 10(2), 107294 (2022).
 13. A. Cydzik-Kwiatkowska, M. Zielińska, "Bacterial communities in full-scale wastewater treatment systems", *World J Microbiol Biotechnol.*, 32(4), 66 (2016).
 14. G. Fernandes Queiroga Moraes, L. V. Cordeiro, F. P. de Andrade Júnior, "Main laboratory methods used for the isolation and identification of *Staphylococcus* spp", *Rev Colomb Cienc Quím Farm.*, 50(1), 5-28 (2021).
 15. T. Sibanda, R. Selvarajan, M. Tekere, "Targeted 16S rRNA amplicon analysis reveals the diversity of bacterial communities in carwash effluents", *Int Microbiol.*, 22(2), 181-189 (2019).
 16. J. Sun, K. Liu, Y. Gao, L. Jin, Y. Gu, W. Wang, "Isolation, plant colonization potential and phenanthrene degradation performance of the endophytic bacterium *Pseudomonas* sp", Ph6-gfp *Sci Rep.*, 4(1), 5462 (2014).
 17. M. Ahmed, A. Elramalli, S. Amri, Y. Abouzeed, "Cefoxitin mannitol salt agar for selective isolation of methicillin-resistant *Staphylococcus aureus*", *Ibnosina Journal of Medicine and Biomedical Sciences.*, 6(1), 31-33 (2014).
 18. B. Jung, G. J. Hoilat, (MacConkey medium), *In: StatPearls (Internet): StatPearls Publishing Treasure Island (FL)*, (2021).
 19. M. Bhuiya, M. K. Sarkar, M. H. Sohag, H. Ali, C. K. Roy, L. Akther, *et al.*, "Enumerating antibiotic susceptibility patterns of *Pseudomonas aeruginosa* isolated from different sources in Dhaka City", *Open Microbiol J.*, 12, 172 (2018).
 20. S. Li, J. Liu, M. Sun, W. Ling, X. Zhu, "Isolation, characterization, and degradation performance of the 17 β -estradiol-degrading bacterium *Novosphingobium* sp. E2S", *Int J Environ Res Public Health.*, 14(2), 115 (2017).
 21. P. Thakur, C. Nayyar, V. Tak, K. Saigal, "Mannitol-fermenting and tube coagulase-negative staphylococcal isolates: unraveling the diagnostic dilemma", *J Lab Physicians.*, 9(1), 65-66 (2017).
 22. A. G. Nikezić, S. Z. Blagojević, M. Đ. Čupurdija, N. S. Planojević, J. V. Jovankić, J. D. Rakobradović, *et al.*, "Comparative analysis of human DNA extraction methods and mitochondrial DNA HV1 and HV2 haplogroup determination", *Kragujevac J Sci.*, (42), 73-83 (2020).
 23. H. K. Byers, E. Stackebrandt, C. Hayward, L. L. Blackall, "Molecular investigation of a microbial mat associated with the great artesian basin", *FEMS Microbiol. Ecol.*, 25(4), 391-403 (1998).
 24. A. Naqib, S. Poggi, S. J. Green, "Deconstructing the Polymerase Chain Reaction II: an improved workflow and

- effects on artifact formation and primer degeneracy", *PeerJ*, 7, e7121 (2019).
25. S. K. Abd-El-Aal, A. Attallah, N. M. Abdel-Aziz, B. E. Khalil, "Isolation, screening, and molecular identification of pectinase producers from fruits, vegetables, and soil samples", *Egyptian Pharm J*, 21(3), 302-311 (2022).
 26. R. Xu, M. Sun, J. Liu, H. Wang, X. Li, X. Zhu, *et al.*, "Isolation, characteristics, and performance of a diethylstilbestrol-degrading bacteria strain *Serratia sp*", *Huan Jing Ke Xue*, 35(8), 3169-3174 (2014).
 27. H. K. Allen, D. O. Bayles, T. Looft, J. Trachsel, B. E. Bass, D. P. Alt, *et al.*, "Pipeline for amplifying and analyzing amplicons of the V1–V3 region of the 16S rRNA gene", *BMC Research Notes*, 9(1), 380 (2016).
 28. H. Jiang, J. Bao, J. Liu, Y. Chen, C. Feng, X. Li, *et al.*, "Development of a Quantitative PCR Method for Specific and Quantitative Detection of *Enterocytopora artemiae*, a Microsporidian Parasite of Chinese Grass Shrimp [*Palaemonetes sinensis*]", *Front Mar Sci*, 8, 730569 (2021).
 29. R. Tian, B. Imanian, "ASAP 2: a pipeline and web server to analyze marker gene amplicon sequencing data automatically and consistently", *BMC Bioinform*, 23(1), 27 (2022).
 30. J. G. Caporaso, J. Kuczynski, J. Stombaugh, K. Bittinger, F. D. Bushman, E. K. Costello, *et al.*, "QIIME allows analysis of high-throughput community sequencing data", *Nat Methods*, 7(5), 335-336 (2010).
 31. B. J. Callahan, P. J. McMurdie, M. J. Rosen, A. W. Han, A. J. Johnson and S. P. Holmes, "DADA2: High-resolution sample inference from Illumina amplicon dat", *Nat methods*, 13(7), 581-583 (2016).
 32. C. Quast, E. Pruesse, P. Yilmaz, J. Gerken, T. Schweer, P. Yarza, *et al.*, "The SILVA ribosomal RNA gene database project: improved data processing and web-based tools", *Nucleic Acids Res*, 41(Database issue), D590-D596 (2013).
 33. H. J. Gwak and M. Rho, "Data-driven modeling for species-level taxonomic assignment from 16S rRNA: application to human microbiomes", *Front Microbiol*, 11, 570825 (2020).
 34. Z. G. Wei, X. D. Zhang, M. Cao, F Liu, Y. Qian, S. W. Zhang, "Comparison of methods for picking the operational taxonomic units from amplicon sequences", *Front Microbiol*, 12, 644012 (2021).
 35. Y. Cai, Y. Sun, "ESPRIT-Tree: hierarchical clustering analysis of millions of 16S rRNA pyrosequences in quasilinear computational time", *Nucleic Acids Res*, 39(14), e95 (2011).
 36. F. S. Lucas, C. Therial, A. Gonçalves, P. Servais, V. Rocher and J. M. Mouchel, "Variation of raw wastewater microbiological quality in dry and wet weather conditions", *Environ Sci Pollut Res*, 21(8), 5318-5328 (2014).
 37. V. Voolaid, E. Donner, S. Vasileiadis and T. U. Berendonk, "Bacterial diversity and antibiotic resistance genes in wastewater treatment plant influents and effluents", Antimicrobial resistance in wastewater treatment processes, *Trends Microbil*, 157, (2017).
 38. K. Samal, S. Mahapatra and M. H. Ali, "Pharmaceutical wastewater as Emerging Contaminants (EC): Treatment technologies, impact on environment and human health", *Energy Nexus*, 100076 (2022).
 39. S. Begmatov, A. G. Dorofeev, V. V. Kadnikov, A. V. Beletsky, N. V. Pimenov, N. V. Ravin, *et al.*, "The structure of microbial communities of activated sludge of large-scale wastewater treatment plants in the city of Moscow", *Sci Rep*, 12(1), 3458 (2022).
 40. N. A. Zhou, A. C. Lutovsky, G. L. Andaker, H. L. Gough and J. F. Ferguson, "Cultivation and characterization of bacterial isolates capable of degrading pharmaceutical and personal care products for improved removal in activated sludge wastewater treatment", *Biodegradation*, 24(6), 813-827 (2013).
 41. L. Fernández, A. Louvado, V. I. Esteves, N. C. Gomes, A. Almeida and Â. Cunha, "Biodegradation of 17 β -estradiol by bacteria isolated from deep sea sediments in aerobic and anaerobic media", *J Hazard Mater*, 323, 359-66 (2017).

42. T. Mainetti, M. Palmisano, F. Rezzonico, B. Stres, S. Kern and T. H. Smits, "Broad diversity of bacteria degrading 17 β -estradiol-3-sulfate isolated from river sediment and biofilm at a wastewater treatment plant discharge", *Arch Microbiol*, 203(7), 4209-4219 (2021).
43. Q. Zhang, C. Xue, G. Owens and Z. Chen, "Isolation and identification of 17 β -estradiol degrading bacteria and its degradation pathway", *J Hazard Mater*, 423(Pt B), 127185 (2022).
44. R. Khattab, N. Elnwishy, A. Hannora, B. Mattiasson, H. Omran, O. Alharbi, *et al.*, "Biodegradation of 17- β -estradiol in water", *Int J Environ Sci Technol*, 16, 4935-4944 (2019).
45. S. S. Salem, A. El-Fattah, H. M. Abdelbasit and S. Mahgoub, "Isolation and characterization of phenol degrading bacteria from industrial wastewater and sewage water", *Zagazig Journal of Agricultural Research*, 48(2), 443-57 (2021).
46. S. Ibrahim, M. A. El-Liethy, K. Z. Elwakeel, MAE-G. Hasan, A. M. Al Zanaty and M. M. Kamel, "Role of identified bacterial consortium in treatment of Quhafa Wastewater Treatment Plant influent in Fayuom, Egypt", *Environ Monit Assess*, 192(3), 161(2020).
47. L. M. Silva-Bedoya, M. S. Sánchez-Pinzón, G. E. Cadavid-Restrepo and C. X. Moreno-Herrera, "Bacterial community analysis of an industrial wastewater treatment plant in Colombia with screening for lipid-degrading microorganisms", *Microbiol Res*, 192, 313-325 (2016).
48. J. Xu S, Yao, M. Ainiwaer, Y. Hong and Y. Zhang, "Analysis of bacterial community structure of activated sludge from wastewater treatment plants in winter", *BioMed Res Int*, 2018,8278970(2018).
49. D. Shu, Y. He, H. Yue and Q. Wang, "Microbial structures and community functions of anaerobic sludge in six full-scale wastewater treatment plants as revealed by 454 high-throughput pyrosequencing", *Bioresour. Technol*, 186, 163-172 (2015).
50. C. Oluseyi Osunmakinde, R. Selvarajan, B. B. Mamba and T. A. Msagati, "Profiling bacterial diversity and potential pathogens in wastewater treatment plants using high-throughput sequencing analysis", *Microorganisms*, 7(11), 506 (2019).
51. D. Numberger, L. Ganzert, L. Zoccarato, K. Mühldorfer, S. Sauer, H. P. Grossart, *et al.*, "Characterization of bacterial communities in wastewater with enhanced taxonomic resolution by full-length 16S rRNA sequencing", *Sci Rep*, 9(1), 1-14 (2019).
52. N. Xie, L. Zhong, L. Ouyang, W. Xu, Q. Zeng, K. Wang, *et al.*, "Community composition and function of bacteria in activated sludge of municipal wastewater treatment plants", *Water*, 13(6), 852 (2021).
53. M. M. Yosef, A. M. Elrefy and F. A. Elfeky, "Metagenomic Analysis Reveals Microbial Communities in Lake Qarun–Egypt", *J Ecol Eng*, 23(2), 70-76 (2022).
54. W. Dong, Z. Cui, M. Zhao, J. Li, "Seasonal and Spatial Variations of Bacterial Community Structure in the Bailang River Estuary", *J Mar Sci Eng*, 11(4), 825 (2023).
55. A. M. El-Sawah, A. H. Afify, A. M. Abdallah and M. Khedr, "Bacteriological studies on wastewater in some cities in Egypt", (AJSWS) 7(1), 1-13 (2023).
56. H. Tian, J. Liu, T. Feng, H. Li, X. Wu, B. Li, "Assessing the performance and microbial structure of biofilms adhering on aerated membranes for domestic saline sewage treatment", *Rsc Advances*, 7(44), 27198-27205 (2017).
57. O. O. Oyedara, C. J. Ruiz-Amaro, N. Heredia and S. García, "Deciphering Bacterial and Resistome Compositions of Sewage Samples from a Major Wastewater Treatment Plant in Northern Mexico", *A Metagenomics Approach*, (2022).



نشرة العلوم الصيدلانية جامعة أسيوط



تحليل المجتمع البكتيري وقدرته على التحلل الحيوي لمادة ١٧- بيتا إستراديول في محطة مياه الصرف الصحي في أسيوط ، صعيد مصر

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تتم إزالة هرمون الإستروجين فى البول والبراز كمركبات هرمونية قطبية غير فعالة، سواء تم إنتاجها بشكل طبيعي من قبل البشر والحيوانات أو تم إستخدامها كوسيلة للتنظافة الشخصية. أثناء التجميع الأولى وخلال مرحلة معالجة مياه الصرف الصحي، يمكن للإنزيمات البكتيرية تعديل تلك المركبات الى الأشكال النشطة لهرمون الإستروجين .

على الرغم من أنه قد أقتراح أن يتم التخلص من هرمون الإستروجين أثناء معالجة مياه الصرف الصحي فى أغلب الأحيان بواسطة التحلل الحيوى، إلا أن هناك القليل من المعلومات حول البكتريا المسؤولة عن تكسير هرمون الإستروجين. وقد أظهرت التجارب الأولية ان العديد من البكتريا الموجودة فى الأمعاء البشرية والكائنات الحية الدقيقة الموجودة فى الفم قادرة على استخدام ١٧-بيتا إستراديول كمصدر وحيد للكربون. لذا فإن الهدف من هذه الدراسة هو عزل وفحص البكتريا القادرة على تكسير مادة ال١٧-بيتا إستراديول من مياه الصرف الصحي باستخدام طرق العزل التقليدية (الوسط الغذائى المناسب لنمو للبكتريا). وكذلك تحديد التوصيف الكيمائى الحيوى، التوصيف الجزيئى وبناء شجرة النشوء للبكتريا من الجينوم المعزول. علاوة على ذلك، تحديد البكتريا غير المزروعة التى قد تكون موجودة فى عينات مياه الصرف الصحي التى تم جمعها بإستخدام تحليل التنوع البكتيرى عن طريق تحليل تسلسل الجينات من خلال الكشف عن جين (16S rRNA) الريبوسومى. وأخيراً، تقييم الكفاءة التفسيرية للسلاسل المعزولة لمادة ال١٧-بيتا إستراديول .

وقد أجريت هذه الدراسة على عينات من مياه الصرف الصحي لتحليل وتحديد المجتمعات البكتيرية. وأظهرت النتائج وجود ثلاث سلالات بكتيرية سالبة الغرام (ASc1, ASc3, and ASc4) تم عزلها من مياه الصرف الصحي كسلالات قادرة على تكسير مادة ال١٧-بيتا إستراديول واستخدامها كمصدر وحيد للكربون. تم التعرف على السلالات المعزولة عن طريق تحليل تسلسل الجينات من خلال الكشف عن جين (16S rRNA) الريبوسومى ووجد أنها تنتمى الى بورديتيلا سبوتيجينا (Bordetella sputigena)، بيجمنتيفاجا ديجنس (Pigmentiphaga daeguensis) ، ستينوتروفوموناس بافيانى (Stenotrophomonas pavaini) على التوالى.

وباستخدام تحليل التنوع البكتيري، تم إكتشاف ٢٨ فئة بكتيرية (bacterial phylum)، ٥٣ طبقة بكتيرية (bacterial class)، ١٠٤ ترتيب بكتيري (bacterial order)، ٢٢٦ عائلة بكتيرية (bacterial family) و ٥٥٩ جنس بكتيري. (bacterial genus) كانت الفئات البكتيرية التالية، (Bacteroidetes, Firmicutes and Proteobacteria) والتي تمثل (٥٦ - ١٨%) من التسلسلات المصنفة هي الأكثر فئات وفرة في جميع التسلسلات. وايضا، تم إكتشاف الفئات البكتيرية التالية، (Actinobacteria, Cyanobacteria, Chloroflexi, Caldithrix and Chlamydiae) والتي تمثل (٠,٠٠١ - ١,١%) من التسلسلات المصنفة هي الأقل فئات وفرة في جميع التسلسلات. فيما يخص الطبقات البكتيرية، أظهرت النتائج ان الطبقات البكتيرية التالية (Bacteroidia, Flavobacteria and Sphingobacteria) هي الأكثر الطبقات وفرة في الفئة البكتيرية (Bacteroidetes phylum) حيث أن (Bacteroidia) تمثل (٤٩,٧%) من جميع التسلسلات المصنفة. فيما يتعلق بالفئة البكتيرية (Firmicutes phylum)، أظهرت النتائج وجود ثلاث طبقات رئيسية داخل هذه الفئة وهي (Clostridia, Bacilli and Erysipelotrichi) حيث تمثل ال 50.1% (Clostridia) من جميع التسلسلات المصنفة والتي تعتبر الأعلى وفرة، في حين تمثل ال 5.1% (Bacilli) من جميع التسلسلات المصنفة. فيما يتعلق بالفئة البكتيرية (Proteobacteria)، تنتمي غالبية التسلسلات البكتيرية المصنفة الى الطبقات البكتيرية التالية، (Gammaproteobacteria, Epsilonproteobacteria, Betaproteobacteria, Deltaproteobacteria and Alphaproteobacteria).

وفيما يتعلق بكفاءة السلالات المعزولة على تكسير مادة ال ١٧-بيتا إسترايول واستخدامها كمصدر وحيد للكربون، كان معدل التكسير للسلالات ستينوتروفوموناس بافياني، بورديتيلا سبوتيجينا و بيجمنتيفاجا ديجنس هو ٦٦,٤%، ٦٥,٥% و ٥٩% على التوالي.